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# 3.3 Estimation of the traffic volume in roadways that were not surveyed

# 3.3.1 Estimation of daily traffic volume

The daytime traffic volume of roadways not subject to monitoring was estimated by applying those described in "BMRT, 1984; Average Daily Traffic". However there were some roadways whose traffic volumes did not allow such approximation and thus required extrapolation based on known traffic volumes of roadways interlinked. The estimation method is shown in Table 3-4.

# 3.3.2 Estimation of traffic volume ratios by time period and by vehicle type

The traffic volume ratio of the roadways not surveyed were assumed to be the same as those of roadways in the same links and in nearly equal road width that were surveyed. The names of roadway assumed are shown in Table 3-4. The traffic volume of (4) Others was monitored jointly about LPG cars (Taxi) and motorcycles. Thus to figure out the ratio of both vehicles, the data (ratio) reported by "JICA, 1986; Study on road improvement, rehabilitation and traffic safety in Bangkok" were reviewed and the ratio of taxi to motorcycle was determined to be 36 by 64, corresponding to the observation within Bangkok city limit.

## 3.3.3 Drive speed estimation

The drive speed of traffic not surveyed was thought same as that of roadways in same link and in similar width that were surveyed. However there were some roads that did not allow this approach and thus drive speed in such roads was assumed unanimously 40 km/hr for those with good surface and 30 km/hr for ones poorly surfaced. As for speedways, the speed was assumed to be 80 km/hr, the speed limit. The Table 3-4 lists the drive speed of motorways applied.

#### 3.4 Determination of emission factors

#### 3.4.1 NO<sub>x</sub> emission factor

The NO<sub>x</sub> emission volume from driving car varies depending upon various traveling conditions such as vehicle type, speed, manufacturing age, and even car manufacturer. In accordance, NO<sub>x</sub> emission factor has to be determined through drive patterning test involving combined effects of the extent and frequency of acceleration, deceleration, and constant navigation. Once drive pattern established, the test run on Chassis-dynamometer will allow the measurement of NO<sub>x</sub> emission volume. However the cost and time incurred forbid such drive patterning and Chassis-dynamometer test. Thus referenced for emission factor determination are data reported by Ministry of Construction, Japan set for six types of vehicles and with respect to varying drive speeds. Since the emission control is not practiced in Thailand, the factors applied are those of non-regulated cars in Japan before 1973 as shown in Table 3-5.

Table 3-4 Summary of traffic flow volume

Name of route   Dieset    Caso   Roler   Total   Caso   Total		Ki See			Daily	traffic	Dally traffic volume for each vehicle types	or each	vehicle		Ave-		Applied	how] and	
1   1   1   1   1   1   1   1   1   1		ź 		_	20		-ose		Woter-		Se s		each vehi-	3 2	
1   1   1   1   1   1   1   1   1   1				,	2	5	Ę	 목	٥	Total	(¥		cle types	sbeed	
2   3   3   3   3   3   3   3   3   3	-				;		3		,				7		_
3   3   2   SIGNAMITT ED.   186805   13314   22991   8722   15567   186805   4.4   17-106-1.)   18-106-1.    17-106-1.	$\exists$		Ξ	STAGE E	2315	83	8721	73	သ	33.55	8				
3   3   4   7   100000   13314   2529   13714   13529   13717   13529   13714   13529   13714   13529   13714   13529   13714   13529   13714   1371	Н		2		16386	13314	882	87728	15517	76336	47	T-I(MS-1) is applied.	T-1(KS-1)	T-1(NS-1)	
3   3   4	_	_	Ш	*	16386	13314	22891	8728	15517	7838	* 47		*T-1(MS-1)	*T-106-1)	
3   3   6	H		-	*	10008	8133	14043	233	9478	48865	47	1984-ADT	T-1(AS-1)	T-I 04S-1)	
3   3   6	H			*	2587	2102	3828	1378	2450	12146	47	(46995 - 34849)	1-108-1)	1-1(85-1)	
3   34   1 DAVIN-TRAT IMP,   58221   71826   5626	-	L	ļ		833	2523	5051	4673	888	28200	9	(17054 + 12146)	KS-4-2		
3   35   4   6   1   1   1   1   1   1   1   1   1	L	L	-1	P	288	1158	88	83	1410	31867	* 88		* AS 5	* 18-5	
3   323 9   SULCAMANT RIAD.   31366   122239   122239   23249   2 45	H	1	œ	*	6221	7182	5658	83	8	20455	æ	J is applied.	Γ.	-8 <del>7</del> -5	
6   3115   10   PICTOMO, SAMING, PILA1   10.	-	_	6		13185	12280	12839	6289	11145	55719	*		-	# T-3	
6   3115   11   11   11   11   11   11	L.	~-	10	PILAÍ	9129	7848	5307	3547	838	32137	* 33			± 1-2	
6   3115   12   PINTINAWCSA RO.   4556   1552   2556   27759   4652   17155   2556   2775   357   35	$\vdash$		5 11	,	10358	6593	1259	4094	1277	34849	× 45			* KS-4-1	
6   3116   13   PHITTINWANCA RO.   1857   716   1158   2511   850   4787   330   8904   4787   13804-MIT   15 septiled.   1457   1569   1560	⊢	-	5 12		9357	1532	988	82.12	4852	17054	98*		HS-4-2	# KS-4-2	
6   3117   14   14   15   15   15   15   15   15	H		Ω		1521	202	:183	ন্ত	88	4787	8	f BWRT 1984-ADT ; is applied.	HS-4-2		
6   3255   15   15   15   15   15   15   15	L		7		DLD CLG	219	88	स	22	1487	8	[ BART 1984-AUT ; is applied.	-12-4-2		
6   3258   15   THEPHMAN RD.   7739   3573   5685   2539   4515   24257   8 30   30   30   30   30   30   30	H		6 15		1180	538	83	88	229	3836	30	f DWRT 1984-AUT ; is applied,	5.		
6   3344   17   11   11   11   11   11   11	Н	_	9	<b>JEPINRAK</b>	7739	3578	5886	2539	4515	24267	* 30		* T-5	* 1-5	
6   3344   18   10176   5077   12135   2843   4689   34690   4.0   465 - 0.0   4.0   4.5 - 0.0   4.5	_		11 17		3351	1501	3539	733	1,08	10333	8		T-4	T-4	
6   33.44   19   10.176   5.077   12.135   25.42   4659   34500   440   4450   440   4450   44	ا		-		10176	5007	12135	2643	883	34860	± 40		7-1 ≠	* I-4	•
6   4001   25   25   25   25   25   25   25   2	-		_		10176	5007	121351	2843	4699	34880			I-4	T-4	
6   4002   21   21   22   705   222   705   222   705   225   40   Balf of @   225   225   40   Balf of @   225   225   40   Balf of @   225	H				5410	8239	0257	443	1410	18411		() - SS#	5-5		
6   4003   22   13   14   12   15   15   15   15   15   15   15		-	_		5012	3114	2460	23	305	320		Half of @	5.5		_
6   4004   22   TIANBAN RD.   5289   3887   3383   2880   3716   17795   30   40-40 (34849 - 17054)   1754   1757   1757   1878   30   40-40 (34849 - 17054)   1757   17	_	_	8		2705	3114	2460	222	705	9206		Half of @	<del>KS-5</del>		
6   4005   24   PARATA PHANAP RD.   136   117   719   525   478   30   11°s assumed 40% of 12°s traffic count use ferry,   1-2     6   4007   55   PARATA PHANAP RIMAN		_	g	THAIBAN RD.	5236	3367	33.3	2330	37.16	17795	R	<b>0-0</b> (34849 - 17054)	MS-4-1		
6   4006   55   WANDAY KNUMN KNAN KNAN KNAN KNAN KNAN KNAN KNAN KN	Н	9	ষ	PARAYA PIKYAP RD.	138	117	. 79	প্র	83	478	ଞ	It's assumed 40% of T2's traffic count use ferry.	1-2	T-2	
6   4007   25   25   25   25   25   25   25   2	-4	6.400	থ	KILLAN		1401	88	83	1127	5733	ଞ	It's assumed that $@: \otimes: \otimes -5: \otimes: 35$ .	r-2	T-2	
6 4008   27   ST   COUNTY		_			815	701	474	316	3	0 83 10 10 10 10 10 10 10 10 10 10 10 10 10	8	half of ♠	1-2	T-2	
5   4008   50   PIETOWILING RD.   551   817   553   359   567   30   1817 of \$400   77   184   329   1817 of \$400   77   1818   329   1817 of \$400   77   1818   329   1817 of \$400   77   72   72   72   72   72   72		-	12		815	701	474	316	ফ্র	2870	8	Half of ♦	T-2	1-2	
6   4010   25   25   27   184   223   1874   30   laif of 69   7-2   7-2   25   4011   30   257   184   252   1874   30   laif of 69   7-2   7-2   25   4011   31   25   25   25   25   25   25   25   2			83	PIETUMING RD.	ૹ	817	553	88	657	337			T-2	1-2	
6 4011 30   476 408 277 184 329 1874 30 Half of 69   1-2		6 4010	8		475	88	112	젎	g	1674		lbif of @	1-2	1-2	
6   4012 31   23   24   25   25   25   25   25   25   25	4	-			476	83	22	182	ខ្ល	1574	_	Half of @	1-2	1-2	
6 3109 32 NA XLOM ND.   1		_	33		88	졅	S.3	ક્ષ્	646	- 83	ଞ	Estimated by T2's traffic count and ferry's count.	T-2	1-2	
6 3012 33 SUKSIMIT-SAMITISMORY RD. 105 105 105 105 105 105 105 105 105 105		_	ਲ									it is need and became traffic flow in a			_
6   2001   34   301 WITHWINGWING   4967   2297   3785   1630   2698   15577   30   (6×6+0)   1-5			ಣ	TSAKION								at a line of the state of the s			
6   3238   55   TIEDYWAKK RD.	님	$\overline{}$	Š	SOI WATDANSAMBONG											_
		<u> </u>	X3	THEPHARAK RD.	4367	1833	3785	1830	2838	15877	88	@×®+@	1-5	1-5-T	

(Note 2) [\*] is observed value.

2 : Expressway 3 : National road 6 : Provincial road

Kind of road

(Note 1)

BMRT; Bangkok Metropolitan Regional Transport Sector Study, Office of the National Economic and Social Development Board.

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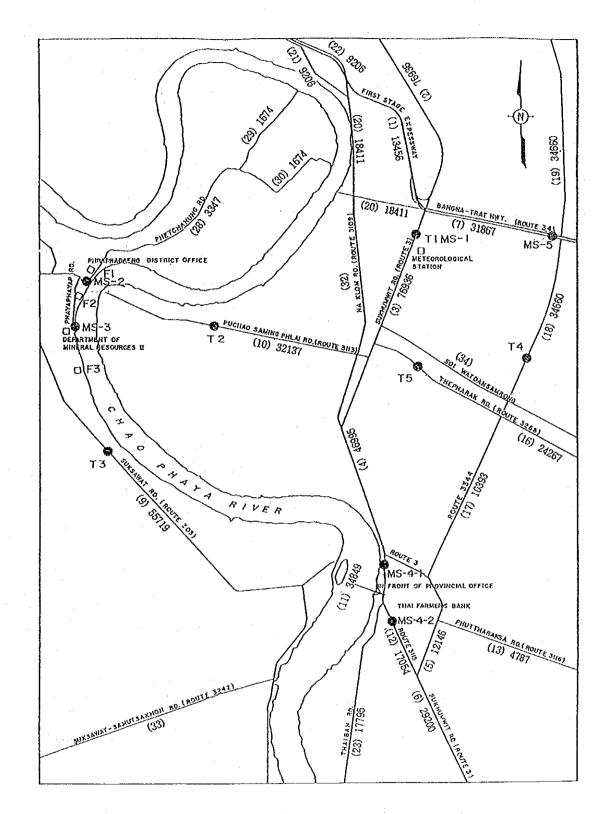


Fig. 3-5 The road map

Table 3-5 NO<sub>x</sub> emission factor by Ministry of Construction in Japan (Non restrictions)

vehicle type	10km/h	20km/h	30km₂/h	40km/h	50km/h	60km/h	80km/h	100km/h	unit
Passenger car	2.62	2.54	2.82	3.15	3.48	3.75	4.65	4.63	:
Light cargo vehicle	6.98	5.59	5.26	5.04	4.74	4.27	4.98	8.03	g/kma/car
Heavy duty gasoline vehicle	7.25	6.04	5.92	5.92	5.84	5.60	7.54	8.68	
Diesel vehicle (Direct fuel injection)	2.074	1.654	1.550	1.478	1.381	1.237	1.143	1.864	
Diesel vehicle (Non-direct fuel injection)	1.185	1.019	1.027	1.057	1.075	1.067	1.489	1.933	g/km/ton
Diesel vehicle (Swirl chamber type)	0.889	0.776	0.792	0.825	0.850	0.855	1.148	1.489	

Table 3-5 lists 6 types of vehicles but monitored in Thailand were 5 types of cars. Furthermore, there were a slight difference observed between the definition of vehicle type in Japan and that of Thailand, that required the additional coordination/matching effort.

## (1) Gasoline car

The factor of non-regulatory car reported by Ministry of Construction, Japan was applied for this type of car in Thailand as shown by Table 3-6.

#### (2) LPG driven car

The 10-mode test value of non-regulated LPG car in Japan (average speed 17.7 km, 3.13 g/km. car) was multiplied by emission volume ratio by drive speed described in Table 3-5 for  $NO_x$  emission factor estimation of this type. The result is shown in Table 3-6.

#### (3) Cargo truck

The cargo truck in Thailand was assumed to be 75 pct diesel car and 25 pct gasoline car. The emission factor (g/km, ton, car) of light cargo was approximated from that of diesel car reported by Ministry of Construction, Japan multiplied by the average weight of car, 1.5 ton (Table 3-7) to determine the value for Light weight vehicle in Thailand. The result is shown in Table 3-6.

The emission factor of Heavy vehicle assumed to be of Diesel type and likewise estimated by using the factor of diesel car (g/km, ton, car) reported by Ministry of Construction, Japan multiplied by the average weight of 9.75 ton (Table 3-7). The ratio of engine types was assumed to be 10%: 45%: 45% for direct injection type, non-direct fuel injection type and swirl chamber type respectively and the factor determined corresponds to the weighted average of this combination. The result is shown in Table 3-6.

#### (4) Motorcycle

The cylinder volume of motorcycle in Thailand is about 125 ml. Thus the factor was assumed to be 1/10 of that of passenger car with the cylinder volume of 1200-1500 ml. The result is also as shown in Table 3-6.

Table 3-6 NO<sub>x</sub> emission factor applied for this study

(unit; g/km/car)

vehicle type	10km/h	20km/h	30kma/h	40km/h	50km/h	60km/h	80km/h	100km/h
Light vehicle	2,75	2.27	2.21	2.19	2.14	2.03	2.54	3.68
Heavy vehicle	11.12	9.49	9.49	9.70	9.79	9.64	12.68	16.83
Gasoline	2.62	2.54	2.82	3.15	3.48	3.75	4.65	4.63
LPG	3.21	3.11	3.45	3.85	4.26	4.59	5.69	5.66
Hotor cycle	0.26	0.25	0.28	0.32	0.35	0.38	0.47	0.46

Table 3-7 Average weight of trucks

(unit: ton)

Vehicle	type	Empty	Loaded	Applied value in this study
Light vehicle	4 wheel	1	2	1.5
11	6 wheel	5	12	8.5 lave.
Heavy vehicle	10 wheel	8	18~25	14.75 9.75 6 wheel: 10 wheel=8:2

## 3.4.2 SO<sub>2</sub> emission factor

The SO<sub>2</sub> emission factor was estimated by the following equation.

SO<sub>2</sub> emission factor (g/km)=(diesel truck rate/fuel consumption(km/
$$\ell$$
))

×specific gravity(kg/ $\ell$ )

×Sulphur content(%)×64/32×10 ......(3-1)

#### Where

Fuel consumption is those described in "Ministry of International Trade and Industry; Manual of Ambient  $SO_x$  and  $NO_x$  prediction Method in Comprehensive Environmental Assessment, 1982" and Japan Environment Agency; Manual of Total Emission Control for sulphur oxides, 1975", which are 3.3 km/liter for a large truck and 5.9 km/liter for a small truck, specific gravity is thought 0.83 for diesel oil, sulphur content was the mean value (0.46%), which was based on the analysis results of the Diesel oil (0.65%, 0.46%) used in Thailand.

The obtained  $SO_2$  emission factors were listed in Table 3-8.

Table 3-8 SO<sub>2</sub> emission factor applied for this study

vehicle type	Emission factor (g/km/car)
Light vehicle	0.97
Heavy vehicle	2.31

#### 3.5 Calculation of SO<sub>2</sub> and NO<sub>x</sub> emission volume

The  $SO_2$  and  $NO_x$  emission volumes (per day and per km) were calculated for each of 31 links by multiplying the emission factor (g/km, car) of each car and its traffic volume (car/hr) for 24 hour period and by summing up the products about all vehicle types.

The Table 3-9 shows day average emission volume of  $NO_x$  and  $SO_2$  per km and total emission volume in each district of Samut Prakarn in Table 3-10. The source distribution figure of  $SO_2$  and  $NO_x$  emission volume from roadway is shown in Fig. 3-6 and Fig. 3-7.

The value is the product of day average of emission volume, total road length in the area and 365 days.

(It would be better to account for an additional 1 km allowance per each link which intersects the provincial borderline.) Hourly variation of emission volume by roadway in whole area is shown in Table 3-11.

Table 3-9 Estimated emission volume per unit length for each route

,	······································		· · ·	
ł			Daily a	
1 .:	Route	length		n volusie
Link				t length
Į I		(ka)	(Na	/b/km)
number	Whole	In Sanut		
	area.	Prakasn	802	NOx
[ <del></del>	5.30	0	0.075	1.318
$-\frac{1}{2}$	4.69	0	0.680	5.773
3	3.31	2.01	0.680	5,773
4	5.28	5.28	0.416	3.527
5	3.77	3.77	0.107	0.911
6	33.88	33.06	0,209	1.638
7	12.07	4.07	0.513	3.580
8	25.38	24.87	0.329	2.298
9	20.27	16.47	0.600	4.270
10	6.40	6.40	0.394	2.509
11	1.10	1.10	0.369	2.585
12	2.06	2.06	0.303	1.013
13	9.69	9.69	0.045	0.311
14				
15	9,48 24,27	9,48 23,32	0.014	0.097
16	12.27	12.27	0.230	0,236 1,577
17			0.230	0.735
18	5.80	5.80 3.58	0.094	
	4.59			2.450
19	5.01	0	0.313	2.450
20	3.97		0.286	1.825
21	2.07	0	0.143	0.912
22	2.02	0	0.143	0.912
23	4.18	4.18	0.188	1.244
24	1.26	1.26	0.006	0.037
25	0.96	0.96	0.070	0.448
	0.88	0.45	0.035	0.224
27	3.79	2.23	0,035	0.224
28	4.79	4.79	0.041	0.261
29	2.22	2.22	0.020	0.131
30	2,82	2.82	0.020	0.131
31	0.91	0.91	0.040	0.257
35	18.50	18.50	0.148	1.012
<u> îotal</u>	242.99	201.55		

Note) (Whole area) includes about 1 km outside area from boundary of Samut Prakarn.

Table 3-10 Estimated emission volume by automobile

County	N	Emission volu	me (ton/year)
code	Name of county	80,	NOx
1	Muang	577.76	3127.45
2	Bang Plee	381.57	1923.84
3	Phra Pradaeng	141.78	686.00
Outo	f Samut Prakarn	372.96	2074.31
	TOTAL	1474.07	7811.60

Table 3-11 Hourly variation of emission volume by road/way in whole area

(unit Nm3/h)

73 X X X X	Enissio	n volume
TIME	SO <sub>2</sub>	NOx
$0:00 \sim 1:00$	16.9	135.7
1:00 ~ 2:00	13.6	98.7
2:00 ~ 3:00	12.9	88.1
$3:00 \sim 4:00$	15.1	98.5
$4:00 \sim 5:00$	21.4	135.1
$5:00 \sim 6:00$	36.3	229.6
6:00 ~ 7:00	58.9	424.3
7:00 ~ 8:00	63.3_	576.9
8:00 ~ 9:00	66.7	528.5
9:00 ~ 10:00	86.7	599.4
$10:00 \sim 11:00$	102.4	711.0
$11:00 \sim 12:00$	102.7	738.6
$12:00 \sim 13:00$	94.6	656.6
$13:00 \sim 14:00$	90.6	636.0
$14:00 \sim 15:00$	96.7	671.2
$15:00 \sim 16:00$	103.1	718.0
16:00 ~ 17:00	88.8	647.6
<u>17:00 ~ 18:00</u>	78.2	649.9
18:00 ~ 19:00	67.9	536.0
$19:00 \sim 20:00$	58.5	448.8
$20:00 \sim 21:00$	50.4	378.7
$21:00 \sim 22:00$	37.4	299.7
$22:00 \sim 23:00$	28.9	242.3
$23:00 \sim 24:00$	21.4	171.1
AVE	58.9	434.2

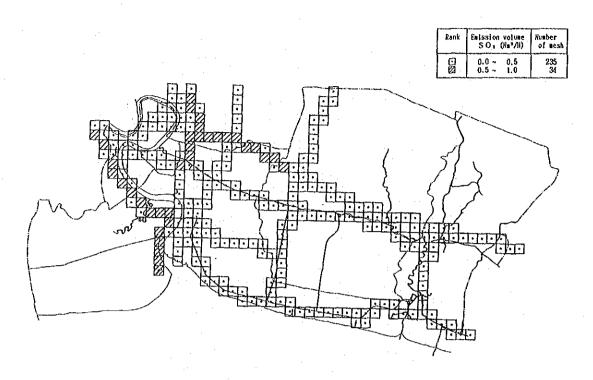


Fig. 3-6 The source distribution of SO<sub>2</sub> emission volume from roadway

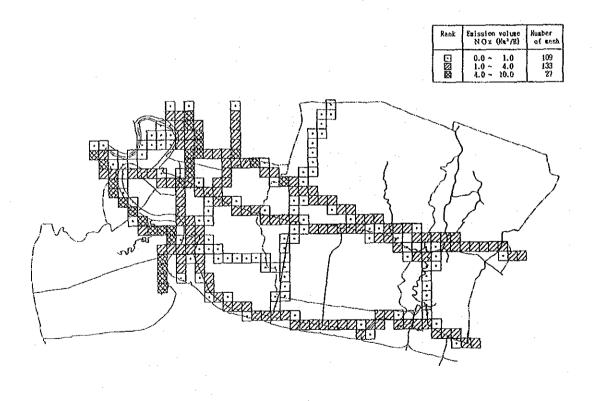


Fig. 3-7 The source distribution of NO<sub>x</sub> emission volume from roadway

## 4. Estimation of SO<sub>2</sub> and NO<sub>x</sub> emission volume from vessels and ferryboats

#### 4.1 General

Ships up- and down-streaming along Chao Phraya river and ferryboats in Phra Pradaeng provincial part of the river are thought to have some environmental impact and thus the SO<sub>2</sub> and NO<sub>x</sub> emission from these sources were estimated.

The  $SO_2$  and  $NO_x$  emission volumes from vessels and ferryboats were estimated (as shown in Figure 4-1) by knowing such variables as number of vessels, horse power of their engines,  $SO_2$  and  $NO_x$  emission factors.

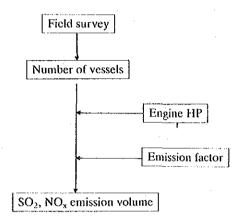


Figure 4-1 Evaluation procedure of SO<sub>2</sub> and NO<sub>x</sub> emission volumes from vessels as well as ferryboats

#### 4.1.1 Ships

As aforementioned, the ferrying frequency was checked in the daytime but not at night. Accordingly, the estimation of nighttime frequency had to resort to the information supplied by Thai counterpart.

Same is true for the tonnage variation of ships on sail along Chao Phraya. The density of boats was assumed to be identical everywhere along the Chao Phraya.

The  $SO_2$  as well as  $NO_x$  volume was calculated about 37.5 km length of Chao Phraya within the Samut Prakarn district and about the additional 1 km of upstream portion and 5 km off shore of the river mouth to the south. It was assumed that ships with 1,000 tonnage or more used fuel oil, and that 50 percent of ships with less than 1,000 tons used high speed diesel oil and other 50 percent used low speed diesel oil.

The main and backup engine of ships and loading factor of them were assumed to be same as those generally in use in Japan.

#### 4.1.2 Ferryboats

The ferryboats stay at the point on the extrapolated line of Puchao Saming Phrai Road, shuttle across the Chao Phraya carrying some 11-14 cars. There are three routes shown as Fig. 4-2, from

the northmost F1 to the South, F2 and F3 and only F1 route is open at midnight. The service is offered on flextime table such that the boat leaves as it is loaded with cargos. The frequency of ferry services at night was not monitored and so it was estimated by using the traffic volume of Puchao Saming Phrai Road.

There are two types of ferryboat, one with 4 engines and the other with two engines suitable for small ferryboat and two engines are run during navigation as well as anchorage. However, the load factor of engines are different between navigation period and anchorage period, thus leads to a difference of  $SO_2$  and  $NO_x$  emission volume. The sulphur content of the high speed diesel oil used for ferryboats was thought same as that of high speed diesel oil commonly in use in Thailand. The horse power and navigation speed of them were monitored through field surveys and then the regular calculation methods popularly known in Japan were used to evaluate  $SO_2$  and  $NO_x$  emission volumes.

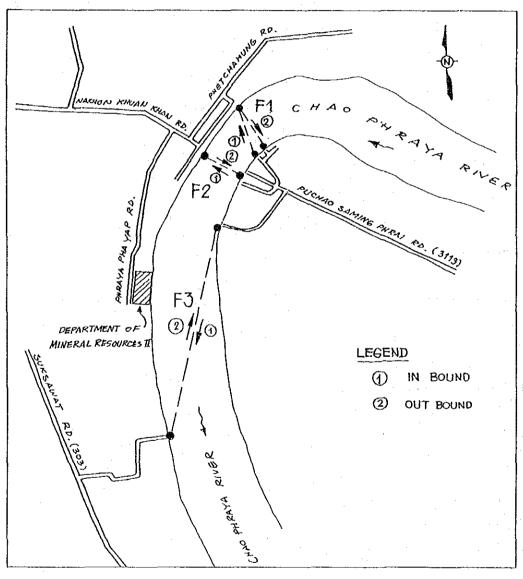


Fig. 4-2 Survey points of vessel and ferryboat navigation

## 4.2 Field survey

The number of boats and ferries in the river was surveyed twice on January 13th and July 13th, 1988 both during 6:00 through 18:00 time frame. The survey points are shown in Fig. 4-2 and the result of this investigations in Table 4-1 and Table 4-2.

Table 4-1 Vessels up/down streaming observed

Time	1st survey 1988/1/13	3rd survey 1988/7/13
$6:00 \sim 7:00$	5	11
$7:00 \sim 8:00$	8	· 10
$8:00 \sim 9:00$	10	10
$9:00 \sim 10:00$	8	4
$10:00 \sim 11:00$	3	12
$11:00 \sim 12:00$	13	4
$12:00 \sim 13:00$	8	11
$13:00 \sim 14:00$	9	11
$14:00 \sim 15:00$	10	3
$15:00 \sim 16:00$	9	4
$16:00 \sim 17:00$	7	13
$17:00 \sim 18:00$	6	4
TOTAL	96	67

Table 4-2 Ferryboats shuttling across the river

Time	1st survey 1988/1/13			3rd survey 1988/7/13		
TIME	F1	F2	F3	F1	F2_	F3
$6:00 \sim 7:00$	49	8	6	46	8	15
$7.00 \sim 8.00$	37	8_	16	39	7_	19_
$8:00 \sim 9:00$	41	8	18	_37_	8_	19
$9:00 \sim 10:00$	52	12	22_	46	12	23_
$10.00 \sim 11.00$	51	14	20	48	12_	20
$11:00 \sim 12:00$	66	_16	- 20	45_	10	_ 22
12:00 ~13:00	_ 58	16_	19	48	10	22
13:00 ~14:00	49	11	18	_52_	15_	30
14:00 ~15:00	52	14	_ 21	_52_	12	34
$15:00 \sim 16:00$	_53_	_13_	15_	51_	8	34_
$16:00 \sim 17:00$	49	12	18	42	_12	25
$17:00 \sim 18:00$	47	12_	11	37_	7_	19_
TOTAL	604	144	204	543	121	282

## 4.3 Estimation of traffic volume at night

#### 4.3.1 Ships

The traffic volume of boats in Chao Phraya river was monitored during daytime from 6:00 through 18:00 and not at night. Thus it was estimated by means of a material available from Harbour Department, "Communication Section, Harbour Department; Incoming and Outgoing Trips of Vessel in Chao Phraya River, 1988".

The material deals with the traffic volume of all vessels of above 1000 tons passing through the river mouth of Chao Phraya with respect to ship name, nationality, ship's agency and passage time. An example is shown in Table 4-3 about January case of 1988.

Anyhow, the daytime (12 hrs) traffic volume was assumed to be the mean value of two on-site survey results and that of night time was estimated by means of traffic pattern described by the material abovementioned and the daytime survey result. Table 4-4 summarizes the number of ships estimated by time period. The seasonal fluctuation of the traffic was disregarded.

Table 4-3 Vessels incoming and outgoing the river mouth of Chao Phraya (over 1000 tonnage, Jan. 1988)

(unit;ship/hour)

Time	Incomm- ing to BKK Port	Outgoing from BKK Port	Total
$0:00 \sim 1:00$	11	li	22
1:00 ~ 2:00	12	28	40
$2:00 \sim 3:00$	11	19	30
$3:00 \sim 4:00$	8	13	21
4:00 ~ 5:00	16	10	26
5:00 ~ G:00	38	8	46
$6:00 \sim 7:00$	49	17	66
$7:00 \sim 8:00$	34	22	56
$8:00 \sim 9:00$	26	27	53
$9:00 \sim 10:00$	22	32	54
$10:00 \sim 11:00$	17	24	41
$11:00 \sim 12:00$	12	25	37
$12:00 \sim 13:00$	16	12	28
$13:00 \sim 14:00$	12	30	42
$14:00 \sim 15:00$	32	33	65
15:00 ~16:00	39	37	76
$16:00 \sim 17:00$	33	39	72
$17:00 \sim 18:00$	46	39	85
$18:00 \sim 19:00$	44	48	92
$19:00 \sim 20:00$	26	38	64
$20:00 \sim 21:00$	37	29	66
$21:00 \sim 22:00$	21	24	45
22:00 ~23:00	24	27	51
23:00 ~24:00	32	30	62
TOTAL	618	622	1240

( unit ; ship/hour )

Time	Number of ships
$0:00 \sim 1:00$	3
$1:00 \sim 2:00$	5
$2:00 \sim 3:00$	4
$3:00 \sim 4:00$	2 3 6
$4:00 \sim 5:00$	3
$5:00 \sim 6:00$	6
$6:00 \sim 7:00$	3
$7:00 \sim 8:00$	9
$8:00 \sim 9:00$	10
$9:00 \sim 10:00$	6
$10:00 \sim 11:00$	7.5
$11:00 \sim 12:00$	8.5
$12:00 \sim 13:00$	4.5
$13:00 \sim 14:00$	5
$14:00 \sim 15:00$	6.5
$15:00 \sim 16:00$	6.5
$16:00 \sim 17:00$	10
$17:00 \sim 18:00$	5
18:00 ~19:00	11
19:00 ~20:00	- 8
20:00 ~21:00	88
21:00 ~22:00	5
22:00 ~23:00	6
23:00 ~24:00	7
TOTAL	149.5

#### 4.3.2 Ferryboats

The shuttling frequency of ferries was also determined by averaging the data taken in these two time surveys during 6:00 through 18:00 period. The number of ferries at night (18:00-6:00) was estimated based on the traffic volume of Puchao Saming Phrai road on T-2 link that leads to the pier of ferryboats and the daytime traffic pattern recorded.

The ratios of ferrying services at night (18:00-21:00) via three routes, F1, F2 and F3 were assumed to be 60.4:14.0:25.6 same as those of daytime. However during 21:00 through 6:00 only F1 route is thought in service.

Table 4-5 summarizes the number of ferries in service across the Chao Phraya in both direction for 24 hours period. The monthly fluctuation of traffic volume was disregarded.

Table 4-5 Ferryboats shuttling across Chao Phraya

(unit : ferrys/hour)

m:	]	Ferry	rout	e
T ine	F 1	F 2	F 3	Total
$0:00 \sim 1:00$	15	0	0	15
$1:00 \sim 2:00$	8	0	0	8
$2:00 \sim 3:00$	6	0	0	6
$3:00 \sim 4:00$	8	0	0	8
$4:00 \sim 5:00$	11	0	00	11
5:00 ~ 6:00	19	0	0	19
$6:00 \sim 7:00$	47.5	8	10.5	66
7:00 ~ 8:00	38	7.5	17.5	63
$8:00 \sim 9:00$	39	8	18.5	65.5
9:00 ~10:00	49	12	22.5	83.5
$10:00 \sim 11:00$	49.5	13	20	82.5
11:00 ~12:00	55.5	13	21	89.5
$12:00 \sim 13:00$	53	13	20.5	86.5
13:00 ~14:00	50.5	13	24	87.5
14:00 ~15:00	52	13	27.5	92.5
$15:00 \sim 16:00$	52	10.5	24.5	87
16:00 ~17:00	45.5	12	21.5	79
17:00 ~18:00	42	9.5	15	66.5
$18:00 \sim 19:00$	39	9	17	65
19:00 ~20:00	41	9 7	17	67
20:00 ~21:00	30		13	50
21:00 ~22:00	29	0	0	29
22:00 ~23:00	28	0	00	28
23:00 ~24:00	24	0	0	24
TOTAL	831.5	157.5	290	1279

#### 4.4 Estimation of SO<sub>2</sub> and NO<sub>x</sub> emission volume

Needless to say, the emission volume of  $SO_2$  as well as  $NO_x$  varies depending on the type of vessel, displacement tonnage, dead weight tonnage and the status of vessel whether it is on sail or at anchor. Accordingly the emission volume is estimated in terms of those variables.

#### 4.4.1 Vessels

#### (1) Size variation of vessels

The traffic volume of vessels in Chao Phraya is approximated roughly 150 per day as stated before but they spread widely in vessel size and thus in term of emission volume. The size variation of vessels is assumed as follows.

a) Vessels of less than 1000 and above 1000 gross tonnage

As shown by the aforementioned material published by Harbour Department, the number of vessels above 1000 tonnage during 12 hours period (6:00-18:00) in June 1986 reports 675 in total and 26 per day as averaged over 26 days of survey period.

On the other hand, the number of vessels monitored at site for 12 hours daytime period was 81.5 per day. Thus the ratio of vessels between above 1000 and less than 1000 gross tonnage is calculated as 26:55.5.

## b) Breakdown of vessels above 1000 gross tonnage

Harbour Department of Thailand keeps the record of vessels at anchor in Bangkok Harbour with respect to vessel name, harbour entering date, and net cargo tonnage. ("SHIPS MOVEMENT and SHIPS ANCHORED AND ETA, Bangkok Bar" by Harbour Department).

The breakdown of vessels above 1000 gross tonnage is estimated with an assumption that the net cargo (displacement) tonnage is nearly equal to gross tonnage.

Listed in Table 4-6 is the estimated breakdown of vessels at anchor in Bangkok Harbour.

GT rank (ton) Ratio (%)

Table 4-6 Breakdown of vessels at anchor in Bangkok Harbour

GT rank (ton)	Ratio (%)		
0 ~ 999	68.1		
1000 ~ 2999	13.6		
3000 ~ 5999	14.2		
6000 ~ 9999	3.7		
10000 ~	0.4		

IV - 60

## (2) Estimation of SO<sub>2</sub> and NO<sub>x</sub> emission volume

The equation used for estimation are those reported on "MITI, Manual of Ambient SO<sub>2</sub> and NO<sub>2</sub> Prediction Method in Comprehensive Environmental Assessment, IPCAJ (1988)".

- a) When rating power output×load factor < 1000ps
  Fuel consumption (kg/Vessel.hr)=0.165×rating output×load factor
  SO<sub>2</sub> emission vol (kg/vessel.hr)=fuel consumption (kg/vessel.hr)×Sulphur (%)
  ×(1/100)×(64/32)

  NO<sub>x</sub> emission vol (Nm³/vessel.hr)=0.00135×(rating output×load factor)

  Exhaust gas vol (Nm³/vessel.hr)=5.46×rating output×load factor

  (4.1)
- b) When rating power output×load factor  $\ge 1000 \text{ps}$ Fuel consumption (kg/vessel.hr)=0.15×rating output×load factor + 25.24 SO<sub>2</sub> emission vol (kg/vessel.hr)=fuel consumption (kg/vessel.hr)×Sulphur (%) ×(1/100)×(64/32) NO<sub>x</sub> emission vol (Nm³/vessel.hr)=0.00182×(rating output×load factor)<sup>1.12</sup> Exhaust gas vol (Nm³/vessel.hr)=5.05×rating output×load factor + 490.96

The rating output and load factor applied in these equations are listed in Table 4-7, same as those presented in the manual.

In regard to sulphur content in fuel, it was assumed that ships with 1,000 tons or greater used fuel oil, so that the mean value (2.5%) of the sulphur content analyses for the four types of fuel oil used in Thailand were applied. In regard to ships displacing less than 1,000 tons, the percentage of ships which use high speed engines and which use low speed engines are about the same, and they use high speed diesel oil and low speed diesel oil, respectively. Therefore, with respect to high speed diesel oil, the mean value (0.56%) of the sulphur content analyses for the two types of diesel oil used in Thailand was used. Regarding low speed diesel oil, the upper limit (1.5%) of the sulphur content standards for marine diesel oil established by the Thai Department of Commerce were used.

The navigation speed of vessels is the average of data monitored at site, 10.8 knot for those of less than 1,000 tonnage and 8 knot for large vessels over 1,000 tonnage. The vessel speed of 8 knot corresponds to the legal limit and is thought reasonable for bulky ones.

The emission factor for vessels of less than 1,000 tonnage is discretely defined for three categories, 0–99 ton, 100–499 ton and 500–999 ton range in Japan. However, the breakdown of vessels in Thailand is not known and so the emission factor is determined with respect to the 250 ton vessel as representative one.

Table 4-7 Horsepower , load factor, sailing speed and fuel sulphur content by vessel size

O.Tti	Output	(HP)	Sulfer	Load	factor	Sal-
GT rank (ton)	main engine	sub englne	tents (%)	main englne	sub engine	Speed (knot)
0 ~ 999	450	40	0.56 1.5	0.85	0.39	10.8
1000 ~ 2999	2,400	131	2.5	0.85	0.39	8
3000 ~ 5999	4,500	232	2.5	0.67	0.39	8
6000 ~ 9999	8,000	337	2.5	0.40	0.39	8
10000 ~	11,825	568	2.5	0.40	0.39	8

## 4.4.2 Ferryboats

The  $NO_x$  and  $SO_x$  volume estimation relies on the equation (4-1) as previously shown and is calculated by applying the values listed in Table 4-8.

Table 4-8 Horsepower, load factor, sailing speed and fuel sulphur content of ferrys

	Term Set up value		Set up base		
Number of engines 2 engin		2 engines	Large ferry has 4 engines. Small ferry has 2 engines. Large and small ferry also normally operate 2 engines.		
			As a result of field survey, it is set up. At F1 and F2 ferry route, average horsepower of 2/(10ferrys) and 200ps(6ferrys) is 225ps.		
Sulfu	r contents	0.56 %	It is averaged value by sulfur contents of 2 kinds of high speed diesel oil which is consumed in Thailand.		
Load	Anchoring	0.5			
factor	Sailing	1.0	As a result of field survey, it is set up.		
Ancho	ring time	2.5 min	As a result of field survey, it is set up.		
Sailing speed		3.72 km∕h	As a result of field survey, it takes about 5 minutes at a distance of 310 meters. Therefore, salling speed is 3.72 km/h. Sailing length is F1;350m, F2;310m, F3;1500m. Sailing time is calculated that sailing length devide by sailing speed.		

## 4.5 The summary of SO<sub>2</sub> and NO<sub>x</sub> emission estimation for ferryboats and ohter vessels combined

Both  $NO_x$  and  $SO_2$  emission volumes are listed for vessels of varying size and ferryboats in Table 4-9 and Table 4-10, respectively. Hourly variation of emission volume by vessels and ferryboats is shown in Table 4-11. Source distribution figure of  $SO_2$  and  $NO_x$  emission volume from vessels and ferryboats is shown in Fig. 4-3 and Fig. 4-4.

Table 4-9 SO<sub>2</sub> and NO<sub>x</sub> emission by vessel size

0.55	SO <sub>2</sub> (ton/year)			NO <sub>×</sub> (ton/year)		
GT rank (ton)	main engine	sub engine	total	main engine	sub engine	total
0 ~ 999	108.7	4.4	113.1	243.6	5.8	249.4
1000 ~ 2999	373.3	9.5	382.8	428.9	6.2	435.1
3000 ~ 5999	561.9	17.6	579.5	693.6	12.7	706.3
6000 ~ 9999	154.9	6.6	161.5	193.2	5.7	198.9
10000 ~	24.4	1.2	25.6	32.3	1.0	33.3
TOTAL	1223.2	39.3	1262.5	1591.6	31.4	1623.0

Table 4-10  $SO_2$  and  $NO_x$  emission from ferryboats

(unit : ton/year)

Ferry	SO <sub>2</sub> emission volume			NOx emission volume		
route	Ancho- ring	Sail- ing	Total	Ancho- ring	Sall- ing	Total
F 1	5.26	23.75	29.01	17.61	89.46	107.07
F 2	1.00	3.98	4.98	3.33	15.01	18.34
F 3	1.63	31.55	33.18	5.35	116.50	121.85
Total	7.89	59.28	67.17	26.29	220.97	247.26

Table 4-11 Hourly variation of emission volume by vessels and ferryboats

(unit Nm3/h)

C 1 1 ( P	Ves	sels	Ferry	boats	TOT	AL
TIME	S O 2	ΝОх	S 0 2	NOx	SO <sub>2</sub>	NOx
0:00 ~ 1:00	24.3	43.5	0.5	2.6	24.8	46.1
1:00 ~ 2:00	40.5	72.4	0.3	1.4	40.8	73.8
2:00 ~ 3:00	32.4	57.9	0.2	1.0	32.6	58.9
3:00 ~ 4:00	16.2	29.0	0.3	1.4	16.5	30.4
$4:00 \sim 5:00$	24.3	43.5	0.3	1.9	24.6	45.4
$5:00 \sim 6:00$	48.6	86.9	0.6	3.3	49.2	90.2
$6:00 \sim 7:00$	24.3	43.5	3.0	15.3	27.3	58.8
$7:00 \sim 8:00$	72.9	130.4	3.4	17.5	76.3	147,9
$8:00 \sim 9:00$	81.0	144.8	3.6	18.3	84.6	163.1
$9:00 \sim 10:00$	48.6	86.9	4.5	22.9	53.1	109.8
$10:00 \sim 11:00$	60.7	108.6	4.2	21.7	64.9	130.3
$11:00 \sim 12:00$	68.8	123.1	4.6	23,3	73.4	146.4
$12:00 \sim 13:00$	36.4	65.2	4.4	22.6	40.8	87.8
$13:00 \sim 14:00$	40.5	72.4	4.7	24.2	45.2	96.6
$14:00 \sim 15:00$	52.6	94.1	5.2	26.4	57.8	120.5
$15:00 \sim 16:00$	52.6	94.1	4.7	24.3	57.3	118.4
$16:00 \sim 17:00$	81.0	144.8	4.2	21.7	85.2	166.5
$17:00 \sim 18:00$	40.5	72.4	3.3	17.1	43.8	89.5
18:00 ~ 19:00	89.1	159.3	3.4	17.6	92.5	176.9
$19:00 \sim 20:00$	64.8	115.9	3.5	18,0	68.3	133.9
20:00 ~ 21:00	64.8	115.9	2.6	13.5	67.4	129.4
$21:00 \sim 22:00$	40.5	72.4	0.9	5.0	41.4	77.4
$22:00 \sim 23:00$	48.6	86.9	0.9	4.8	49.5	91.7
$23:00 \sim 24:00$	56.7	101.4	0.8	4.1	57.5	105.5
AVE	50.4	90.2	2.7	13.7	53.1	103,9

Rank	Emission volume S.O. (Na³/ii)	Number of mesh
	0.0 ~ 1.0 1.0 ~ 2.0 2.0 ~ 3.0	31 20 1

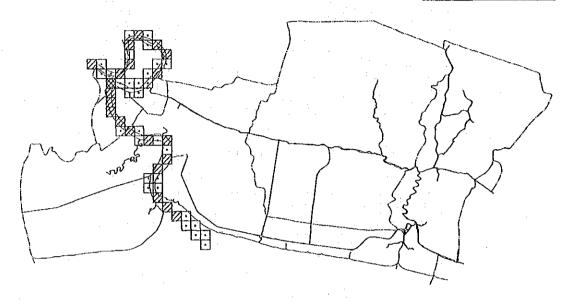


Fig. 4-3 Source distribution of SO<sub>2</sub> emission volume from vessels and ferryboats

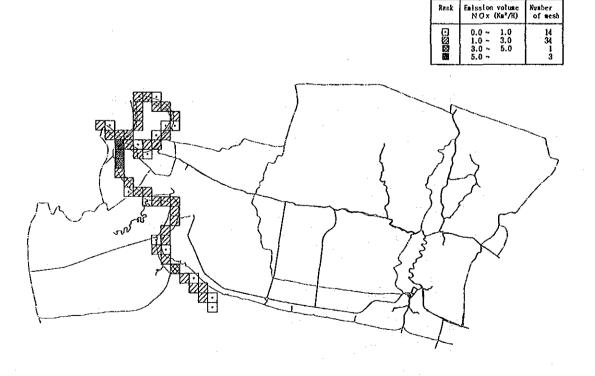


Fig. 4-4 Source distribution of  $NO_x$  emission volume from vessels and ferryboats

# 5. $SO_2$ and $NO_x$ emission volume emitted from the total area of Samut Prakarn

The total emission volumes, as summed up for those coming from all sources such as factories, automobiles, boats and ferries, are as shown in Table 5-1 and these pollutant emissions by each mesh segment area in Figure 5-1 and Figure 5-2.

As shown by them, the total  $SO_2$  emission from this specified are break into 86.7% from the factories (18,330 tons/y), 7.0% from automobiles (1,474 tons/y) and 6.3% from vessels and ferries (1,764 tons/y). Samely the total  $NO_x$  volume is shared by factories 47.7% (8,820 tons/y), by automobiles 42.2% (7,812 tons/y) and by vessels and ferries 10.1% (1,870 tons/y).

Table 5-1 Summary of emission volume in 1988

				ii
Name of source		Type of source	SO <sub>2</sub> emission volume (ton/year)	NO <sub>x</sub> emission volume (ton/year)
Questionnaire return		point	13649	8108
Stationary sources	Questionnaire nothing	area	4681	712
	Sub total		18330	8820
Ro	ad way	line	1474	7812
	Vessels (sailing)	point	1263	1623
Vessels and Ferryboats	Ferryboats (anchoring)	point	8	26
	Ferryboats (sailing)	point	59	221
	Sub tota	al	1330	1870
	TOTAL		21134	18502

Rank	Emission volume S O . (Nm³/H)	Humber of mesh
	0.0 ~ 1.0 1.0 ~ 10.0 10.0 ~ 100.0 100.0 ~	364 90 18

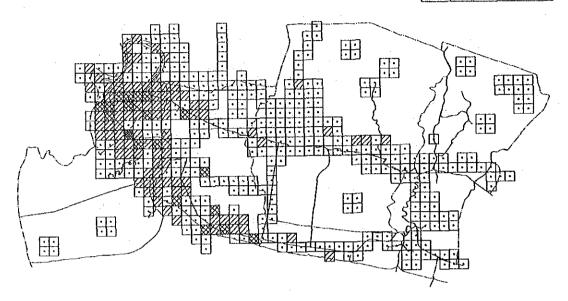


Fig. 5-1 Source distribution of SO<sub>2</sub> emission volume from all sources

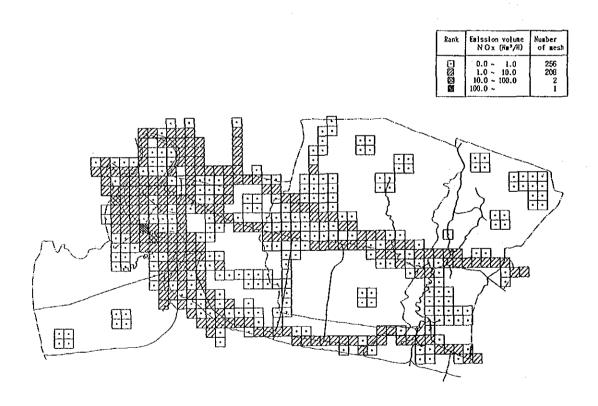


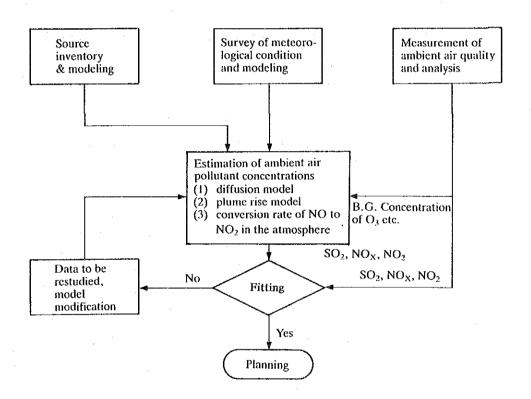
Fig. 5-2 Source distribution of  $NO_x$  emission volume from all sources

PART V ANALYTICAL STUDY ON CURRENT AMBIENT POLLUTANT CONCENTRATIONS BY USING ATMOSPHERIC SIMULATION MODELS

The atmospheric pollutant (SO<sub>2</sub> and NO<sub>2</sub>) concentrations across the Samut Prakarn district is estimated by using simulation models to represent atmospheric diffusion status of the area and the obtained values are subject to analytical studies, comparison with standards and estimation of contribution of each source to atmospheric concentration. As for the particulate matter, the samples collected at the site are analysed and by knowing the chemistry of such dust samples the survey team estimated the contribution of each source by means of chemical mass balance method.

## 1. Atmospheric Simulation Models for Evaluation of Ambient SO2 and NO2 Concentrations

Using such data as SO<sub>2</sub> and NO<sub>x</sub> emission volume from factories, roadways and vessels, meteorological and ambient air pollutant concentration data measured at monitoring stations, atmospheric pollutant concentrations to represent the whole area of Samut Prakarn province as of 1988 are calculated by using the air diffusion model as shown in the following flowchart.



Flow Chart of Simulation Process for SO2, NOx and NO2

#### 1.1 Simulation Methods

#### 1.1.1 Physical and Numerical Simulation Models

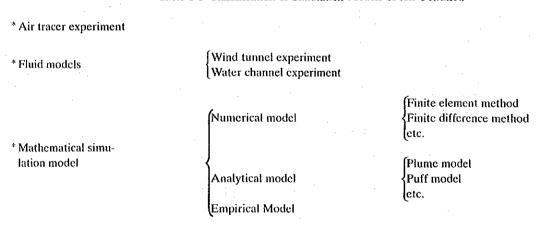
Up until now, several atmospheric diffusion simulation methods have been proposed. These are roughly classified into two groups, the physical laboratory models such as wind tunnel and water channel, and the mathematical models. The classification of those models is given in Table 1-1.

#### (1) Air Tracer Experiment and Laboratory Model Simulation

The air tracer experiment is a direct method of diffusion study, being comprised of such steps as the release of air tracer substances from a site where plant construction is planned, sampling of the diluted gas and analysis of such sample gas. As for the air tracer, SF<sub>6</sub> (Sulfur Hexa Fluoride) is used because the substance is non-toxic for both humans and plants. The air tracer substance also should be easily detectable by a chemical method, even in trace amounts. By analyzing obtained diffusion results together with the meteorological data, one can evaluate detail characterisites of atmospheric diffusion behavior at the construction site.

As for the diffusion over complex terrain or abnormal diffusion problems near stacks involving down-wash or down-draft, a fluid simulation model such as a wind tunnel or a water channel is often employed since the mathematical modelling is prohibitive. In fluid model simulation, the tracer substance is emitted from model stacks and its diffusion is examined visually or by using measuring instruments.

Table 1-1 Classification of Simulation Models of Air Pollution



#### (2) Mathematical Models

Fluid simulation model requires special facilities like a wind tunnel or a water channel and it is not adequate to obtain forecasts of changing meteorological or source conditions. However, mathematical models to represent the diffusion mechanism have flexibility and versatility to evaluate the atmospheric concentrations for changing meteorological and pollutant sourcing conditions. Presently available mathematical models, however, have application limits and difficulties in applying them for special diffusion problems involving complex terrain, buildings, and meteorological conditions. Therefore, we have to select proper models each time depending on the nature of the problems, e.g. a mathematical model for a long term diffusion problem or a fluid model for a terrain affected diffusion.

The pollutants released into the atmosphere are transported and dispersed by wind. If we can assume that the diffusion of pollutant is proportional to the gradient of concentration, then the diffusion is expressed by the following equation. Almost all of the mathematical models adopt this diffusion equation.

$$\frac{\partial C}{\partial t} + u \frac{\partial C}{\partial x} + v \frac{\partial C}{\partial y} + w \frac{\partial C}{\partial z} 
= \frac{\partial}{\partial x} (Kx \frac{\partial C}{\partial x}) + \frac{\partial}{\partial y} (Ky \frac{\partial C}{\partial y}) + \frac{\partial}{\partial z} (Kz \frac{\partial C}{\partial z}) + Q \qquad (1-1)$$

where:

Kx: eddy-diffusion coefficient in the x-direction (m<sup>2</sup>/sec)

Ky: eddy-diffusion coefficient in the y-direction (m<sup>2</sup>/sec)

Kz: eddy-diffusion coefficient in the z-direction (m<sup>2</sup>/sec)

Q: Source intensity (unit/sec)

## 1) Numerical model

Numerical models solve the diffusion equation by the finite difference method or the finite element method using a computer. They are thought effective for unstable conditions and also for complex meteorological conditions. But they often require long computation time and expenses. It is not recommended for use in the prediction of the annual average concentrations for a simulation of long term prediction of average concentration under changing meteorological conditions that incur considerable expense. In accordance, numerical methods are thought suitable for roadway diffusion problems to which analytical models can not be applied or for diffusion of chemically active pollutants such as hydrocarbon and oxidants.

#### 2) Analytical models

An analytical model is one in which an analytical solution to the diffusion equation is used to estimate concentration of pollutants. Many solutions have been propsed for different boundary and flow conditions. Among them, the Gaussian plume and Gaussian puff models are most commonly in use.

#### (1) Gaussian plume model

The Gaussian plume model is the solution of the Fickian type diffusion equation in which pollutants are emitted continuously at a constant flow rate. Then the pollutant disperses as shown by a plume and in concentration whose lateral and vertical profiles are expressed by the Gaussian distributions (Fig. 1-1).

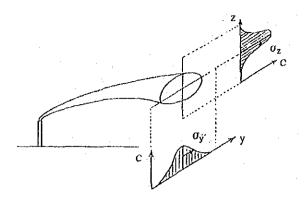


Fig. 1-1 Schematic Representation of the Gaussian Plume Model

In the equation, the ground point where the source stack stands is the origin (0,0,0), the x-axis is the downwind direction and the y-axis is taken normal to the x-axis. The z-axis is the vertical direction. Therefore, the coordinate of the source point is (0,0,He), where He is the effective plume height, i.e., the sum of the stack height and plume rise. (Fig. 1-2).

The distribution of flue gas concentration along y-axis on the plane vertical to the central axis of the plume often complies with the Gaussian distribution profile, in which the concentration maximum lies right on the center axis and then decreases symmetrically as the distance from the axis gets larger. The standard deviation of flue gas concentration,  $\sigma$  is used to indicate the spread of its distribution.

The vertical distribution along the z-axis on the same plane seems to show a somewhat different profile. Neverthless, it can be also approximated by a Gaussian model.

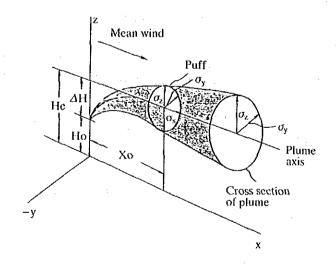


Fig. 1-2 Coordinates of Smoke Plume and Puff Diffusion Model

The concentration of a pollutant at a point (x, y, z) is given by

$$C(x, y, z) = \frac{Q}{2\pi\sigma_{y}\sigma_{z}u} \exp(-\frac{y^{2}}{2\sigma_{y}^{2}}) \left\{ \exp(-\frac{(z-He)^{2}}{2\sigma_{z}^{2}}) + \exp(-\frac{(z+He)^{2}}{2\sigma_{z}^{2}}) \right\} ...(1-2)$$

where symbols are as follows:

C (x, y, z): Concentration of a pollutant at a point (x, y, z)

x, y, z: Longitudinal, lateral and vertical coordinate of the point. x and y are measured from the source point and the lateral distance from the axis which passes through the source point, respectively. z is the vertical height of the point above the ground.

Q: Emission rate of the pollutant  $(m^3/s)$ 

u: Wind velocity (m/s)

He: Effective stack height (m)

 $\sigma_{v}$ : Plume width in lateral direction (m)

 $\sigma_z$ : Vertical plume width (m)

The plume equation (1-2) requires the following conditions in its application.

- The pollutant should be gaseous or particles which are small enough to make falling velocity negligible.
- ii) The pollutants should be inert gases or non-reactive substances which will not cause chemical reactions.
- iii) The pollutants cannot be absorbed or released through the ground surface.
- iv) Wind velocity and direction should be steady in space and time.
- v) The diffusion coefficient is assumed to be identical across the whole target space.
- vi) Emission rate of pollutant is constant and does not change with time.
- vii) The diffusion in the x direction can be negligible as compared with the effect of wind.

These assumptions are not always necessary conditions for us to solve the equation (1-1). In fact, the solution is obtainable without some of above constraints but it takes often a very complex form. An example is the Walter's equation<sup>1)</sup> upheld when the wind velocity and diffusion coefficient change along the height from the ground. If diffusion coefficients Ky, Kz are assumed uniform across the space, one can get the plume equation (1-2) of the Gaussian type even when Ky and Kz vary with time.

According to the statistical theory by Taylor<sup>2)</sup>, the following equation (1-3) is upheld between diffusion coefficient K and diffusion width  $\sigma$  for 2-3 hours from the start of diffusion when the atmospheric turbulency is steady and uniform.

$$\sigma^2 \approx 2Kt$$
 ..... (1-3)

The vertical plume diffusion width is strongly influenced by the atmospheric stability. In the unstable atmospheric condition where fluctuation of wind is prominent in the vertical direction, the plume disperses and  $\sigma_z$  gets larger quickly. On the other hand, in a stable case, (this is the case that air temperature increases as the position gets higher and inversion takes place) turbulence is suppressed and  $\sigma_z$  gets smaller. The plume width depends strongly on the turbulence intensity and several formulae to estimate the plume width in terms of turbulence intensity have been proposed. But from a practical point of view, turbulence data to cover long term period are not easily obtainable. Thus, Pasquill's method which gives vertical and lateral plume widths for each stability category (A~F) in graphic form is broadly applied. Pasquill's stability classification is described in Table III-1-5.

Pasquill's chart for  $\sigma_y$  and  $\sigma_z$  are given in Fig. 1-3. The plume widths are estimated from the figures at any downstream poin of between 0.1 to 100 km and with respect to each stability class.

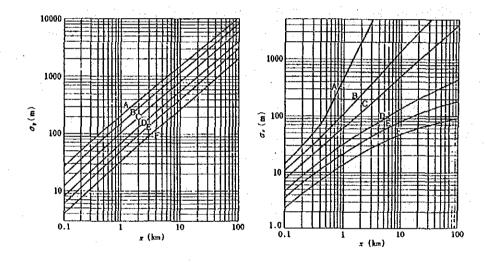


Fig. 1-3 Plume Width by Pasquill-Gifford

This diagram shows the spread of the diffused smoke plume<sup>3)</sup> to the downstream direction at six typical atmospheric stabilities that were first empirically developed based on experiments done with prairie grass and then modified by Gifford.<sup>4)</sup> Turner reported the diagram in his text dealing with air diffusion as the Pasquill-Gifford diagram.<sup>5)</sup> The diffused plume width was originally measured about a smoke source positioned at 0.5 m high on a flat grassland for the period of 10 minutes (later reported by his paper is 3 minutes<sup>6)</sup>). Neverthless, the diagram is effectively applied to the distance range of 100 m to 1 km or to outside of this range by extrapolation. In any case, the Pasquill-Gifford diagram is popularly in use because of its simplicity.

The diffusion width of the plume in the open air is nearly identical both in the vertical and horizontal directions in a few minutes but as time becomes longer than half an hour

(or an hour), the horizontal plume width becomes larger than that of the vertical width due to the nature of air characteristics that the scale of eddy currents are more pronounced in the horizontal than in the vertical direction. The eddy current in the horizontal direction works to add to the vector of plume diffusion as time passes by while the plume width  $\sigma_z$  shows minor change vertically as shown in Fig. 1-4 and the concentration dilutes quickly.

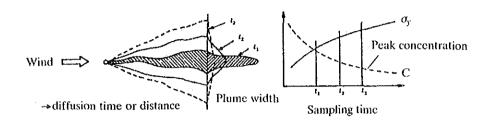


Fig. 1-4 Horizontal Plume Width for Different Sampling Time

The horizontal diffusion width varies with observation time and generally increases proportionally to the power of the observation time ratio  $(t_1/t_2)$ . Thus by applying  $\sigma_y/x\approx\sigma_A$ , the equation (1-4) is upheld.

$$\sigma_{v1}/\sigma_{v2} = \sigma_{A1}/\sigma_{A2} = (t_1/t_2)^m$$
 (1-4)

where;

 $\sigma_{v1}$ : diffusion width at observation time  $t_1$ 

 $\sigma_{A1}$ : standard deviations of plume spread angle at observation time  $t_1$ 

m: power coefficient

The value of the power coefficient is reported to be  $0.2\sim0.3$  by Hanna<sup>6)</sup>. According to a short time model<sup>7)</sup> developed by the Texas Air Control Committee, the diffusion width  $\sigma_{y,t}$  at observation time t, is calculated by knowing the diffusion width  $\sigma_y$ ,  $t_0$  ( $t_0=10$  min) as follows:  $\sigma_{y,t}=\sigma_y$ ,  $t_0/(t_0/t)^m$ 

The value m is a constant given by Table (1-2) for each atmospheric stability group. P-G diagram shows the relationship between  $\sigma_y$  and x for varying atmospheric stability, in which the D group corresponds to a neutral point and the equation  $\sigma_y=0.147x^{0.89}$  is upheld for approximation of the  $\sigma_y$  value (within a 3 minute time range). The equation  $\sigma_y=0.31 \cdot x^{0.89}$  is obtained when the equation  $\sigma_y=0.147x^{0.89}$  is corrected to the hourly equivalent, by using average power coefficient 0.25.

Table 1-2 Time Correction Factor m of Horizontal Plume Diffusion Width

Atmospheric stability group	Α	В	С	D	Е	F
m	0.675	0.55	0.425	0.30	0.175	0.175

## (2) Gaussian puff model

The basic idea of the puff model is to express a continuous plume by a scries of discrete puffs as illustrated in Fig. 1-5. In the Gaussian puff model, the concentration profiles take Gaussian form along the x, y and z axes (Fig. 1-2).

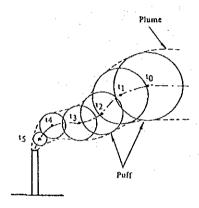


Fig. 1-5 Schematic Representation of Gaussian Puff Model

The concentration profile at time t after the start of emission takes a Gaussian form (1-5) as the solution to the equation (1-1).

$$C = \frac{Q}{(2\pi)^{3/2} \sigma_x \sigma_y \sigma_z} \exp \left\{ -\frac{(x-ut)^2}{2\sigma_x^2} - \frac{(y-vt)^2}{2\sigma_y^2} - \frac{(z-wt)^2}{2\sigma_z^2} \right\} .....(1-5)$$

The equation (1-5) can be expressed as equation (1-6) when the position of the source is assumed at He and the wind blows steadily in the X direction. Both equations (1-5) and (1-6) are thus called the "puff model" for calculation of concentrations.

$$C(x, y, z, t) = \frac{Q}{(2\pi)^{3/2} \sigma_x \sigma_y \sigma_z} \exp\left[-\frac{(x-ut)^2}{2\sigma_x^2} - \frac{y^2}{2\sigma_y^2}\right].$$

$$\left\{ \exp\left[-\frac{(z-He)^2}{2\sigma_z^2}\right] + \exp\left[-\frac{(z+He)^2}{2\sigma_z^2}\right] \right\} ....(1-6)$$

Equation 1-6 expresses the concentration by a puff and the actual concentration from a continuously emitting source is calculated by summing up concentrations of all puffs. Accordingly, the equation is

$$C = \int_{0}^{\infty} \frac{Q}{(2\pi)^{3/2} \sigma_{x} \sigma_{y} \sigma_{z}} \exp \left\{ -\frac{(x-ut)^{2}}{2\sigma_{x}^{2}} - \frac{y^{2}}{2\sigma_{y}^{2}} \right\}$$

$$\left[ \exp \left\{ -\frac{(z-He)^{2}}{2\sigma_{z}^{2}} \right\} + \exp \left\{ -\frac{(z+He)^{2}}{2\sigma_{z}^{2}} \right\} \right] dt \qquad (1-7)$$

The puff models do not require the previously mentioned restrictive conditions (iv), (vi) and (vii) to be applied and are thought suitable for changing the intensity of emission and wind velocity because the concentration profile is estimable by numerical integration. However, they normally involve a considerable volume of computation, though the puff models do not neglect the diffusion to the downstream direction unlike th plume model. In any case, when the same conditions of diffusion are assumed, both equations of the puff model and the plume model give a similar calculation result at wind velocity above 1 m/s. When the wind velocity maintains zero or is negligibly small, the equation (1-8) is derived from the equation (1-7) by introducing u=0 and further by assuming the diffusion widths  $\sigma_x$ ,  $\sigma_y$ , and  $\sigma_z$  being proportional to the lapse of time ( $\sigma_x=\sigma_y=\alpha t$ ;  $\sigma_z=\gamma t$ ), the equation (1-8) is modified into equation (1-9).

$$C = \int_{0}^{\infty} \frac{Q}{(2\pi)^{3/2} \sigma_{x} \sigma_{y} \sigma_{z}} \cdot \exp(-\frac{x^{2}}{2\sigma_{x}^{2}} - \frac{y^{2}}{2\sigma_{y}^{2}}) \cdot \left[\exp\{-\frac{(z - He)^{2}}{2\sigma_{z}^{2}}\} + \exp\{-\frac{(z + He)^{2}}{2\sigma_{z}^{2}}\}\right] dt$$
(1-8)

$$C(R, Z) = \frac{Q}{(2\pi)^{3/2}\gamma} \left\{ \frac{1}{R^2 + \frac{\alpha^2}{\gamma^2} (He - Z)^2} + \frac{1}{R^2 + \frac{\alpha^2}{\gamma^2} (He + Z)^2} \right\} \quad ..... (1-9)$$

where;

R is the horizontal distance between the source and calculation point (m)  $R^2 = x^2 + y^2$ 

In calculation of atmospheric concentrations by the puff model when the wind velocity maintains nearly zero, the diffusion widths  $\sigma_x$ ,  $\sigma_y$  and  $\sigma_z$  of the smoke have to be expressed as a function of the lapse of time (t) as previously explained. The Turner diagram<sup>8)</sup> shown below (Fig. 1-6) is often used to approximate such diffusion widths with some modification to be mentioned later. The Turner diagram is primarily derived from the Pasquill-Gifford diagram while the wind velocity is kept at about 5 m/s and thus it is thought inappropriate to apply the diagram directly to the zero wind case.

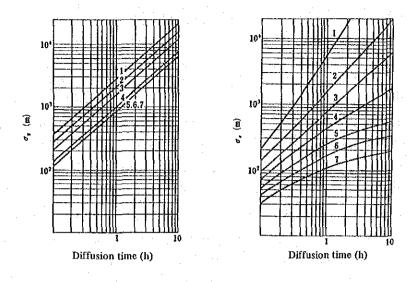


Fig. 1-6 The Diffusion Width by Turner

## 1.1.2 Methods for Calculation of Effective Stack Height

The smoke emitted from the stack mouth reaches a certain height as shown in Fig. (1-7) due to the combined effect of buoyancy coming from the temperature difference between smoke and surrounding air and the upward force vector of smoke itself. The height is called the effective stack height (He) and is the sum of the stack height (Ho) and plume rise ( $\Delta$ H).

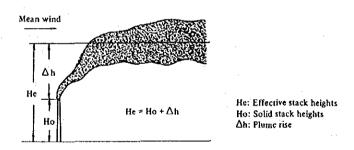


Fig. 1-7 The Relation between Plume Rise AH and Effective Plume Height

In the calculation of plume and puff equations, the pollutant is conventionally assumed to be emitted at the effective stack height. For the plume rise, many formulae, theoretical, empirical or semi emipirical, have been proposed. Table 1-3 shows the representative plume rise equations together with their application conditions.

There are a number of equations proposed to calculate the effective stack height that can be grouped into two types, one of empirical origin and the other theoretically derived. Table 1-3 lists such equations with application specifics. The value  $(\Delta h)$  varies from equation to equation and sometimes gives the difference as large as two times of one to the other at zero wind velocity but the equation by CONCAWE is reported recently as one which offers the best agreement between calculated and observed values.<sup>9)</sup>

Table 1-3 Plume Rise Equation and its Application Range

		Plum rise equation	Theoretical		Basic conditions of used data			
	Year	Q <sub>II</sub> (keal/sec). U (m/sec). H <sub>S</sub> (m). ΔT (°K) T <sub>o</sub> (°K). V (m/sec). Q, (m²/sec). Δθ/Δτ (°K/m)	or empirical	Stability condition	Q <sub>H</sub> (kcal/sec)	11 <sub>S</sub> (m)	U (m/see)	V (m/sec)
l Bosanquet Ist formula	1950	$\begin{split} \Delta h &= \Delta h_m + \Delta h_h \\ \Delta h_m &= [4.77/(1+0.43U/V)] \cdot \sqrt{Q_* \cdot V}/U \\ \Delta h_h &= 6.37g \frac{Q_* \cdot \Delta T}{U^3 \cdot T_a} (1nJ^2 + \frac{2}{J} - 2) \\ J &= \frac{U^2}{\sqrt{Q_* \cdot V}} \cdot (0.43 \sqrt{\frac{T_a}{g\Delta \theta \theta \Delta z}} - 0.28 \frac{v}{g} \cdot \frac{T_a}{\Delta T}) + 1 \end{split}$	Theoretical (Bosanquet)	Unstable to neutral $\Delta \theta / \Delta z = 0.003$	70~1500		3.1~8.1	8~10.8
2 Bosanquet 2nd formula	1957	$\begin{split} \Delta h &= AU[f_1(a) + f_1(a_0) - \{0.615 \alpha_0^{1/2} / (< V/U >^2 + 0.57)^{1/2} \} \} \\ A &= g \cdot Q_{\gamma} \Delta T/2\pi C^2 f_0^{\alpha} U^3 \\ a &= (1 + f_0) / A = (1/A_1) + \alpha_0 \\ a_0 &= f_0 / A - 1 = x/U - t_0 = 4 + V \cdot f_0 / 3g \Delta T \\ t_{10} &= 20 \sec C = 0.13 \\ f_1(a) \cdot f_0(a_0) \end{split}$	Theoretical (Bosanquet)	Unstable to neutral t +1 $t_0 = 200$ see at stable t + $t_0 = 1.572 \times [2T_0/g\Delta\theta/\Delta z]^{1/2}$	70~1500	_	3.1~8.1	8~10.8
3 Bosanquet II modified at CRIEPI	1965	Diffusion coefficient $C=0.13$ of Bosanquet 2nd formula replaced by $C=0.26U^{-1/2}$	Theoretical (Bosanquet)	Same as obove	1500~17000	80~200	I~15	13.2~35
4 CERL formula (Lucas 2nd formula)	1967	∆h ≈ (104+0.17 Hs)O <sub>it</sub> <sup>t/4</sup> /U	Theoretical (Priestley)	Consideration not given	70(N)~25(NX)	76~183	1.5~11.3	14~24.5
5 Moses & Carson formula	1968	$\Delta h = (C_1 VD + C_2 Q_1)^{1/2}/U$ $C_1 = 0.35, C_2 = 5.41$ (at neutral)	Regression equation	C1. C2 different at unstable or stable	2~2500	34~183	l~11.7	2.2~24.5
6 CONCAWE formula	1968	$\Delta h = 5.53 \mathrm{Qz}^{1/2} \mathrm{U}^{-3/4}$	Regression equation	Consideration not given	70~4000		4~10	3.3~17
7 Briggs formula	1969	$\Delta h = 1.6F^{1/3}U^{-1}(3.5x^{K})^{2/3}  x^{K} = 34F^{2/3}(F>55 \text{ m}^{4}/\text{sec}^{2})$ $\Delta h = 5.35 \text{ Q}_{H}^{10.5}/U (\text{Q}_{H} > 1500 \text{ kcal/sec})$	Dimensional analysis (Briggs)	Applied at unstable to neutral, another formula for stable	1000~21000	61~183	1.8~16.5	5.5~28.9
8 TVA formula	1971	$\Delta h = 114 \text{CF}^{1/3}/\text{U}$ , $C = 1.58 - 41.4 \Delta \theta / \Delta z$ $\Delta \theta / \Delta z = 0$ $\Delta h = 60 \text{Q} \pi^{1/3}/\text{U}$	Dimensional analysis	Δθ/Δz>0.003	70X0~250X0	76~183	1.5~11.3	14~24.5
9 Okameto, Okanishi and Shiozawa formula	1979	$\Delta h = 5.5 Q x^{1/2} / U^{3/5}$	Regression equation	Consideration not given	10047000	20~366	1~17	2~40

## 1.2 Development of SO<sub>2</sub>, NO<sub>x</sub> and NO<sub>2</sub> Diffusion Prediction Model

There are various air pollution prediction methods as described in the preceding section. Each of these methods has its own advantages and disadvantages. The method to be used for prediction of air pollution must be determined depending on the purpose of prediction, complexity of computation, the geography of the area and other conditions. The purpose of this study is to calculate the degree of contribution of pollutant sources at present and in the future at Samut Prakarn province in Thailand. In a prediction, it is necessary to calculate the long-term average concentration (annual average) while considering the occurrence frequency of the various meteorological conditions. The region is relatively flat in geography so that the effect of the topography on the diffusion space may safely be ignored.

For these reasons, in this study, the Gaussian plume model and the Gaussian puff model were employed as the diffusion models to calculate the diffusions of all emission sources in Samut Prakarn province in 1988 and to check the agreement between the calculated and actual values measured at stations. Then, parameters such as pollutant emission volume, diffusion parameters, etc. used in this calculation was verified to develop the diffusion prediction models for the SO<sub>2</sub>, NO<sub>x</sub> and NO<sub>2</sub> pollutants. These SO<sub>2</sub>, NO<sub>x</sub> and NO<sub>2</sub> diffusion prediction models consist of the meteorological model, the source model, the diffusion model, the NO<sub>2</sub> conversion model and the annual average prediction model as shown in Fig. 1-8. The components of each model are listed in Table 1-4.

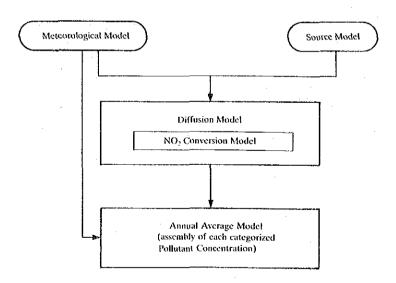


Fig. 1-8 Schematics of SO<sub>2</sub>, NO<sub>x</sub> and NO<sub>2</sub> Prediction Models

Table 1-4 Components of SO<sub>2</sub>, NO<sub>x</sub> and NO<sub>2</sub> Prediction Models

Component	Content	Description			
Meteorological Model	Meteorological frequency	Frequency of occurrence by wind direction, wind velocity and atmospheric stability			
	Upper layer wind velocity equation	Calculation by law of exponent			
	Ground level wind velocity equation	Calculation by law of exponent			
Source Model	Source data	Information such as position, height, emission volume, etc. of each source			
	Plume rise equation	Calculation by CONCAWE equation or Briggs equation			
Diffusion Model	Diffusion equation	Calculation by Gaussian plume equation or Gaussian puff equation			
	Diffusion width	Windy: Yamamoto and Yokoyama's method was employed for factories and ships. The highway model diagram was used for cars.  Calm: Diffusion width under the calm condition by specified by the Ministry of International Trade and Industry was adopted.			
NO <sub>2</sub> Conversion Model	NO <sub>2</sub> conversion equation	Exponent approximation type NO to NO <sub>2</sub> conversion equation			
	Conversion speed	NO to NO <sub>2</sub> conversion rate coefficient			
x 2	O <sub>3</sub> background	Center value by atmospheric stability of ozone concentration measured over ONEB roof was used.			
prediction model		Concentration by meteorological condition was weight-averag accounting for its meteorological frequency and the annual average was calculated.			

## 1.2.1 Meteorological Model

The meteorological conditions governing the atmospheric transport and diffusion were treated as represented by wind direction velocity and atmospheric stability and, thus modelling of the meteorological conditions was made for use in the diffusion calculation as discussed below.

## (1) Grouping of Meteorological Conditions by Season and Time Zone

The seasonal variation and daily variation of meteorological conditions as described in Part III of Section 1.1.1 were divided by seasons and time zones as shown in Table 1-5. Then the agreement between the measured values and the calculated ones was checked with respect to such divisions.

Table 1-5 Grouping of Meteorological Conditions by Season and Time Zone

Season Time	Wet Season May to October	Dry Season November to January	Intermediate February to April
Daytime	7:00 to 17:59	7:00 to 17:59	7:00 to 17:59
Nighttime	18:00 to 6:59	18:00 to 6:59	18:00 to 6:59

#### (2) Division into Meteorological Blocks

The area of Samut Prakarn is nearly uniform in meteorological conditions as stated in Part III of 1.1.5 so that the diffusion may be calculated using the meteorological conditions measured at any of the MS1, MS2 and MS5 stations. When calculating the atmospheric concentration at a mesh point, however, the best reproduction of actual concentration is expected by the use of meteorological conditions measured at the station closest to the mesh point. Therefore, the area was divided into the three meteorological blocks as shown in Fig. 1-9 taking such factors into account as the meteorological correlation among measuring stations, the geography of the area, and the distribution of the pollutant sources, etc. The diffusion calculation was conducted for the two cases, namely, one with MS3 and MS4 (on which meteorological conditions were not monitored) included in the same block of MS1 and the other with MS3 and MS4 included in the block of MS2. It was found that the latter case gave a better agreement between the measured and calculated values. Thus, the division into the meteorological blocks was finally adopted as shown in Fig. 1-9. In the diffusion calculation of ambient pollutants released from factories and ships, the meteorological conditions (occurrence frequency of wind direction, wind velocity and atmospheric stability) of the area where such sources located were applied and in the calculation of the pollutants released from vehicles, the meteorological conditions of the block to which the mesh point belonged were used.

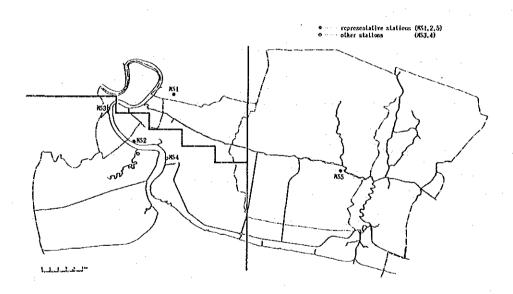


Fig. 1-9 Division into Meteorological Blocks and Representative Stations

#### (3) Wind Direction and Velocity

The wind direction was divided into 16 directions as shown in Fig. 1-10 and the wind velocity into six ranks as shown in Table 1-6, with representative wind velocity established for each rank.

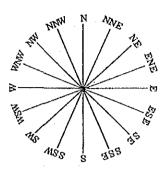


Fig. 1-10 Wind Direction Division

Wind velocity rank Representative wind velocity CALM Calm 0~0.4 m/s 0.7 m/s 0.5~0.9 m/s 1.5 m/s 1.0~1.9 m/s 2.0~2.9 m/s 2.5 m/s Windy 3.0~4.9 m/s 4.0 m/s 5.0~ 6.0 m/s m/s

Table 1-6 Wind Velocity Ranking and Representative Wind Velocity

#### (4) Atmospheric Stability

In this study, applied was the stability classification of the Ministry of International Trade and Industry of Japan which is based on Pasquill's stability classification but uses net radiation flux instead of cloud amount (Table III-1-6). According to this classification, the stability was classified into ranks A to F in windy condition and ranks CA to CD in calm condition. The diffusion width for use in the diffusion calculation was set according to these atmospheric stability ranks (to be discussed later).

# (5) Estimation of Upper Layer and Surface Wind Velocity

In the planetary boundary layer up to an altitude of about 2 km, the wind velocity decreases with decreasing altitude. This is because of the influences of surface roughness and hydrodynamic head loss of atmospheric fluid. This means that smoke released at a lower altitude near the ground surface and smoke released at a higher altitude behave and diffuses differently. In this study, therefore, the air layer was divided into upper, middle and surface layers according to the level of emission sources and the wind velocity in the upper and in surface layers was estimated as listed in Table 1-7.

Table 1-7 Wind Velocity Used in Diffusion Calculation 10)

Layer	Pollution sources	Wind velocity application
Upper	Stationary source with actual stack height 30 m or higher	Wind velocity at actual stack height was estimated by law of exponent.
Middle	Stationary source with actual stack height lower than 30 m and ship	Wind velocity at the representative station was applied as it was.
Surface	Cars	Wind velocity at 3 m above ground was estimated by law of exponent.

In the estimation of wind velocity in the upper layer and ground surface, Equation (1-10) was used. In the equation, the value of index P should ideally be determined by analysis of the data continuously measured in the upper air layer of Samut Prakarn. In this study, however, such data was not available and so the value of P used in the CDA Manual by the EPA of the U.S. (Table 1-8) was adopted.<sup>11)</sup>

$$U_{z} = U_{z0} \left(\frac{Z}{Z_{0}}\right)^{p} \tag{1-10}$$

where;

U<sub>2</sub>: Wind velocity at altitude Z (m/s)

 $U_{z0}$ : Wind velocity at altitude  $Z_0$ , MS1;  $Z_0$ :30 m, MS2 and 5;  $Z_0$ : 10 m

p: Constant that depends on the meteorological conditions (exponent)

Table 1-8 Value of Exponent P

Stability	A	В	С	D	E	F • G
Р	0. 1	0, 15	0, 2	0. 25	0, 25	0.3

Note: These values were reported by De Marrais<sup>11</sup> based on the wind velocity measured on the ground and at the top of a 145 meter high tower located in Brookhaven. The data represents the geography of a gently sloped area near the sea scattered with pine-trees.

#### 1.2.2 Source Model

In order to use the source data collected and consolidated by factories, cars and ships in the diffusion calculation, the sources were modeled as shown below.

#### (1) Smoke Source Type

The sources were classified into the smoke types shown in Table 1-9.

Source Smoke type Factories covered by the Point sources Stationary questionnaire Factories not covered by the Area sources of 2 km by 2 km questionnaire Line sources Cars Ships Vessels (sailing) Point sources Ferries (at anchor) Point sources Ferries (sailing) Point sources

Table 1-9 Classification of Sources for the Diffusion Calculation

# (2) Time Changing Source Activity

For stationary sources, the volume of pollutants released was calculated by season and by time zone (Table 1-5) with respect to the facility operating status and operating time zone (see Part IV of Table 2-9). As for cars and ships, the variations in traffic volume and number of ships by daytime and nighttime determined by the field survey were taken into consideration. The seasonal variation was ignored because of insignificant monthly variations in the traffic volume and in number of ships.

#### (3) Effective Stack Height

1) Effective stack height of stationary (point) source

The effective stack height is the actual stack height  $(H_0)$  plus the plume rise  $(\Delta H)$  of smoke. The following plume rise equations were used.

(1) Windy condition: CONCAWE equation

$$\Delta H = 0.175 \cdot Q_{H}^{\frac{1}{2}} \cdot U^{-\frac{3}{4}} \qquad (1-11)$$

where;

$$Q_H = \varrho \cdot Q \cdot C_P \cdot \Delta T$$
 (cal/s) (released heat)  
 $\varrho = 1.293 \times 10^3 \text{ g/m}^3$  (air density at 0°C)

C<sub>P</sub>=0.24 cal/g.°K (isopiestic specific heat)

Q: Volume of gas emitted (Nm<sup>3</sup>/s)

ΔT: Temperature difference between exhaust gas and atmosphere

$$\Delta T = (T_G - 28)^{\circ}C$$

T<sub>G</sub>: Exhaust gas temperature (°C)

U: Wind velocity at stack top (m/s)

(2) Calm condition: Briggs stable state equation

$$\Delta H = 1.4 \cdot Q_H^{\frac{1}{4}} \cdot (d\theta/dz)^{\frac{3}{8}}$$
 (1-12)

where the following values were used for  $d\theta/dz$ .

Daytime:  $d\theta/dz=0.005$ °C/m (average temperature gradient)

Nighttime:  $d\theta/dz=0.010^{\circ}$ C/m (isothermal layer)

- Effective stack height of stationary source (area source)
   Considering an actual stack height of about 20 m, the effective stack height was set at 30 m.
- Car effective stack height
   The effective stack height of cars was set at 3 m<sup>10</sup>.
- 4) Ship and ferry effective stack height

  The effective stack height of ships was set at 20 m. That of ferries was set at 5 m. <sup>10)</sup>

#### 1.2.3 Diffusion Model

In this study, the Gaussian plume model and the Gaussian puff model were used. The diffusion calculation equations and diffusion width used are described below.

#### (1) Diffusion Equation

The diffusion equations listed in Table 1-10 were applied according to the smoke source types. The diffusion equation is as shown below.

Table 1-10 Sources and Diffusion Equations Used

,	Diffusion equation						
Source	Smoke source type	Windy (u>0.4 m/s)	Calm (u≦0.4 m/s)				
Stationary point source and ship	Point sources	Plume equation {Equation (1-13)}	Puff equation {Equation (1-14)}				
Stationary area source	Area sources	Area smoke source plume equation {Equation (1-15)}	Area smoke source puff equation (Equation (1-16))				
Car	Line sources	Line smoke source plume equation {Equation (1-18)}	Line smoke source puff equation (Equation (1-19))				

# 1) Plume Equation (applied to point sources under the windy condition)

The diffusion concentration by the plume equation can be expressed by the equation (1-13). The ground surface just under the smoke source is set as the origin, and the x-axis is set in the leeward direction from the origin and the y-axis in the horizontal direction at a right to the x-axis. The z-axis is set in the vertical direction.

$$C(x, y, z) = \frac{Q_{P}}{2\pi\sigma_{v}\sigma_{z}u} \cdot \exp\left(-\frac{y^{2}}{2\sigma_{v}^{2}}\right) \left[\exp\left\{-\frac{(z-He)^{2}}{2\sigma_{z}^{2}}\right\} + \exp\left\{-\frac{(z+He)^{2}}{2\sigma_{z}^{2}}\right\}\right] \dots (1-13)$$

where;

x: x coordinate of the calculation point (m)

y: y coordinate of the calculation point (m)

z: z coordinate of the calculation point (m)

Q<sub>p</sub>: Intensity of the point smoke source (m<sup>3</sup>/s)

u: Wind velocity (m/s)

He: Effective stack height (m)

C(x, y, z): Concentration at the calculation point (x, y, z)

 $\sigma_{v}$ : Diffusion width of smoke in y direction (m)

 $\sigma_z$ : Diffusion width of smoke in z direction (m)

#### 2) Puff Equation (applied to point smoke source under the calm condition)

The concentration of atmospheric diffusion by a point sources under the calm condition was calculated by the equation (1-14).

$$C(R, z) = \frac{Q_P}{(2\pi)^{3/2} \gamma} \left\{ \frac{1}{R^2 + \frac{\alpha^2}{\gamma^2} (He - z)^2} + \frac{1}{R^2 + \frac{\alpha^2}{\gamma^2} (He + z)^2} \right\} \dots (1-14)$$

where, R is the horizontal separation (m) between the point smoke source and the calculation point and thus  $R^2=x^2+y^2$  and are the proportional constants of the diffusion width and the transportation time T(s) and thus  $\alpha=\sigma_y/T$  and  $\gamma=\sigma_z/T$ . As the duration of calm condition is at most one hour or so as observed during field surveys, the values of  $\alpha$  and  $\gamma$  at T=3600 sec. were used.

# 3) Plume Equation for Arca Smoke Sources (applied to area sources under the windy condition)

Because the stationary area sources cannot be easily located, it was handled as a 2 km by 2 km area source. The diffusion concentration by the area smoke sources were calculated by subdividing the area source into small  $d\eta \times d\xi$  sources as shown in Fig. 1-11. These small sources were considered as point smoke sources and the concentration was calculated by surface-integration of the equation (1-13).

$$C(x, y, z) = \int_{x-a}^{x+a} \int_{y-a}^{y+a} \frac{Q_A}{2\pi\sigma_y\sigma_z u} \cdot \exp\left(-\frac{\eta^2}{2\sigma_y^2}\right)$$

$$\times \left[\exp\left\{-\frac{(z-He)^2}{2\sigma_z^2}\right\} + \exp\left\{-\frac{(z+He)^2}{2\sigma_z^2}\right\}\right] d\eta d\xi \quad ......(1-15)$$

where;

Q<sub>A</sub>: Release rate per unit area (m<sup>3</sup>/s/m<sup>2</sup>)

a: 1/2 of area source side length (m)

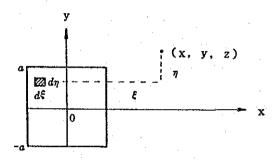


Fig. 1-11 Conceptual View of Surface Smoke Source (windy case)

If the calculation point is within the area source (x < a), the x integration range is x - a to x.

4) Puff Equation for Area Sources (applied to an area source under the calm condition) For convenience of calculation, the diffusion concentration by the area sources was calculated by setting the 2 km by 2 km square area source being equivalent to a circle of the same area, namely a radius a'=1.13 km as shown in Fig. 1-12, subdividing it into small dθ×dr smoke sources, then regarding these small sources as the point smoke sources and integrating the point source puff equation (1-14) as shown below.

$$C(R, z) = \int_0^{a'} \int_0^{2\pi} \frac{Q_A}{(2\pi)^{3/2} \gamma} \left\{ \frac{1}{R'^2 + \frac{\alpha^2}{\gamma^2} (He - z)^2} + \frac{1}{R'^2 + \frac{\alpha^2}{\gamma^2} (He + z)^2} \right\} r dr d\theta$$
 (1-16)

where;

a': Radius of equivalent circular smoke source (m)

R': R' =  $\sqrt{R^2+r^2-2rR\cos\theta}$ 

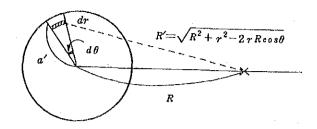


Fig. 1-12 Conceptual View of Area Sources under the Calm Condition

5) Plume Equation for Line Sources (applied to line smoke source under the windy condition)
The ground level concentration by exhaust gas from a line source can be determined by
line-integrating the point source equation, Equation (1-13). Practically, it was calculated by
subdividing the line source into N equal parts as shown in Fig. 1-13, and regarding them as
the point sources and superposing the point source equations.

$$C(x, y) = \sum_{i=1}^{N} \frac{Qi}{2\pi u \sigma_{y} \sigma_{z}} exp\left(-\frac{y_{i}^{2}}{2\sigma_{y}^{2}}\right) \left\{ exp\left(-\frac{(He-z)^{2}}{2\sigma_{z}^{2}}\right) + exp\left(-\frac{(He+z)^{2}}{2\sigma_{z}^{2}}\right) \right\} ... (1-17)$$
 where,

Q: Emission rate of No.i subdivision line source (point source) (m<sup>3</sup>/s)

y<sub>i</sub>: Distance between the calculation point and the x-axis parallel to the wind direction and passing through the center of the No.i subdivision line source (m)

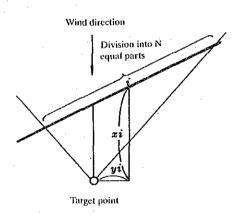


Fig. 1-13 Conceptual View of Atmospheric Diffusion Concentration Calculation by Line Source Diffusion Equation (under the Windy Condition)

6) Puff Equation for Line Sources (applied to line sources under the calm condition)

The atmospheric diffusion concentration under the calm condition from a car line source can be determined by line-integrating the point source puff equation, Equation (1-14). Actually, it was calculated, as such was the case with the calculation of the point source under the calm condition, by letting  $\sigma_y = \alpha T$  and  $\sigma_z = \gamma T$ , subdividing the angle  $\theta$  between the calculation point to the road line into equal angles (Fig. 1-14), substituting the point

sources for the line sources, calculating the diffusion concentrations and superposing these concentrations.

$$C(R) = \sum_{i=1}^{N} \left[ \frac{Qi}{(2\pi)^{3/2} \alpha^2 \gamma} \left\{ \frac{1}{\frac{r_i^2}{\alpha^2} + \frac{(Hc-z)^2}{\gamma^2}} + \frac{1}{\frac{r_i^2}{\alpha^2} + \frac{(Hc+z)^2}{\gamma^2}} \right\} \right] \dots (1-18)$$

where:

- Q<sub>i</sub>: Emission rate from small segment obtained by subdivision of the line source into N equal parts (m<sup>3</sup>/s)
- r<sub>i</sub>: Distance between the calculation point and small segment obtained by subdivision of the line source into N equal parts (m)

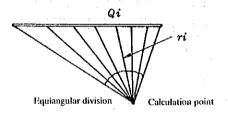


Fig. 1-14 Conceptual View of Atmospheric Diffusion Concentration by Line Source Diffusion Equation (under the calm conditions)

# (2) Diffusion Parameter (Diffusion Width)

For numerical calculation of the ambient air using the diffusion equation, it is necessary to give specific values to the diffusion width in the equation. In this study, the following diffusion widths were used depending on the smoke source height:

Stationary source, ship and ferry under the windy condition ... Yamamoto & Yokoyama's equa-

Diffusion Widths of Stationary Source, Ship and Ferry when with Wind
 The atmospheric diffusion concentrations of the pollutants emitted from the stationary source, ship and ferry under the windy condition were calculated using Yamamoto & Yokoyama's diffusion parameter<sup>12</sup>.

$$\frac{\sigma_{y} = 0.018 \ \sigma_{A} \cdot x}{\sigma_{z} = 0.0054 \ k \sigma_{E} \cdot x}$$
(1-19)

where;  $\sigma_y$  and  $\sigma_z$  are the diffusion widths (m) in the horizontal and vertical directions.  $\sigma_A$  and  $\sigma_E$  are the standard deviations (deg.) in the horizontal and vertical directions. x and k are the constants determined by the leeward distance (m) and the atmospheric stability, respectively. The constant k values are listed in Table 1-11. This table also contains  $\sigma_A$  and  $\sigma_E$  values by the stability calculated from the atmospheric turbulence values measured with

an ultrasonic wind direction and velocity meter. Fig. 1-15 shows the relationship between the leeward distance x and the  $\sigma_y$  and  $\sigma_z$  values by the atmospheric stability calculated from the Equation (1-19) using these values. In comparison with the P-G diagram, the horizontal diffusion width  $\sigma_y$  obtained by Yamamoto/Yokoyama's diffusion width is much larger than that based on the P-G diagram. This is because the P-G diagram represents a three-minute diffusion width while Yamamoto/Yokoyama's represents a one-hour diffusion width.

Table 1-11 Diffusion Parameter K,  $\sigma_A$  and  $\sigma_E$  by Atmospheric Stability Classifications

Parameter	A, B	С	D daytime	D Nighttime	Е	F
K Values	2.67	2.04	1.0	1.0	0.83	0.83
$\sigma_{\Lambda}$ (degree)	26.1	21.6	21.1	20.3	20.0	18.9
σ <sub>E</sub> (degree)	11.7	8.8	7.9	7.5	6.7	6.2

Fig. 1-15 Yamamoto/Yokoyama's Diffusion Width

## 2) Diffusion Widths of Cars under the Windy Condition

The diffusion widths of cars under the windy condition were calculated using the HIWAY diagram of the EPA<sup>13)</sup>. These diffusion widths are expressed by the Equation (1-20) and shown in Fig. 1-16.

$$\sigma_{y} = 465.1(x+b)\tan\theta_{P}$$

$$\theta_{P} = c - d \cdot \ln\left(\frac{x+b}{X_{0}}\right)$$

$$\sigma_{z} = g\left(\frac{x+a}{X_{0}}\right)^{h}$$
(1-20)

where;

x: Leeward distance (km)

a: Coefficient for initial diffusion width  $\sigma_{z0}$  (km)

b: Coefficient for initial diffusion width  $\sigma_{v0}$  (km)

c: Coefficient for  $\theta_p$  (deg.)

d: Coefficient for  $\theta_p$  (deg.)

g: Coefficient for  $\sigma_z$  (m)

h: Exponent for  $\sigma_z$ 

 $\theta_p$ : A half of the angle at which plume varies on the horizontal face

 $X_0$ : =1 km

These coefficient values are listed in Table 1-12. They have been set for  $\sigma_z$  of about 1.5 m at x=0 m,  $\sigma_y$  of about 3 m (a half of a car length) and for agreement with the Pasquill-Gifford diagram.

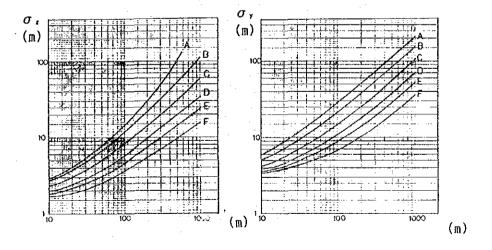


Fig. 1-16 Diffusion Widths in HIWAY Model

In order to improve the consisting between the measured values and calculated values in the actual calculations, the diffusion width was deviated a half rank to the unstable side (e.g., the diffusion range B-C was used in case of Atmospheric stability C rank).

Table 1-12 Coefficients of  $\sigma_0$ ,  $\sigma_2$  used in HIWAY Model

Step of HIWAY	Distance (	km) <sup>Note 1)</sup>	Value for	$\theta_{\rm p}$ (deg.)	Other			
model diagran	am a	ь	c	ď	g (m)	h		
٨	0.00944	0.00863	24.167	2.5 3 3 4	1 2 2.8	0.9447		
В	0.01226	0.0132	18333	1.8096	90.673	0.93198		
c ·	0.01736	0.0210	1 2.5	1.0857	61.141	0.91465		
D	0.02722	0.0348.	8.333	0.72382	34.459	0.86974		
E	0.03590	0.0471	6.25	0.54287	24.26	0.8366		
F	0.05842	0.0733	4.167	0.36191	15.209	0.81558		

Note 1: a and b are the coefficients for the initial diffusion widths  $\sigma_{50}$  of 3.0 m,  $\sigma_{a0}$  of 1.5 m.

#### 3) Diffusion Widths under the Calm Condition

The estimation method of the diffusion widths under the calm condition has not yet been established well. In general, the Turner diagram may be applicable. However, even when there is no wind at about 10 m above the ground, there is wind in the upper layer. Therefore, if the concentrations under the calm condition are calculated using this Turner diagram, they may not agree well with the measured values. The Ministry of International Trade and Industry reported that if the atmospheric diffusion concentrations under the calm condition are calculated by using the diffusion widths several times larger than the values listed in the Turner diagram, they reproduce well the measured values. This diagram is shown in Fig. 1-17. In this study, the atmospheric diffusion concentrations under the calm condition were calculated using this diagram.

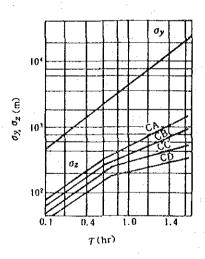


Fig. 1-17 Atmospheric Diffusion Widths under the Calm Condition by the Ministry of International Trade and Industry<sup>10)</sup>

## 4) Initial Diffusion Width $(\sigma_{z0})$

#### (1) Stationary Source (Point Source)

The plume rise  $\Delta H$  changes with the wind velocity so that  $\Delta H$  will change and smoke will disperse vertically near the sources if the wind velocity changes. Therefore, if the plume rise is higher, its variation becomes larger and the initial dispersion width  $\sigma_{z0}$  becomes larger as well. By comparing the TVA diagram<sup>14</sup> developed based on the actual stack data and the diagram obtained from the point source diffusion tests, and also with assumption that the value of  $\sigma_z$  at the short distance value (x=100 m) being equal to the initial diffusion width  $\sigma_{z0}$  caused by plume rise. The initial diffusion width becomes about 25 m under highly stable condition or about 30 m under unstable condition.

Here, the initial diffusion width  $\sigma_{z0}$  of the stationary source (point source) was assumed to increase with the plume rise  $\Delta H$ , 30 m under unstable and neutral conditions at  $\Delta H$ =100 m and 25 m when it was stable. The value  $\sigma_{z0}$  does increase with increasing  $\Delta H$  but by taking the data on which the TVA diagram originates and other factors into consideration, it was assumed that the initial diffusion width would reach its upper limit value at  $\Delta H$ =100 m and it would not increase further even if  $\Delta H$  becomes higher. The lower limit value was set at 5 m. The assumed relationship is illustrated in Fig. 1-18.

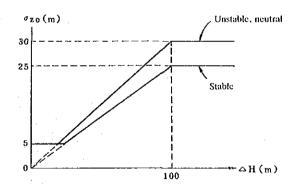


Fig. 1-18 Initial Diffusion Width  $\sigma_{z0}$  of Stationary Source (Point Source)<sup>10)</sup>

② Stationary Source (Area Sources), Car, Ship and Ferry
The initial diffusion width  $\sigma_{z0}$  of the stationary source (area source), car, ship and ferry was set at 5 m<sup>10)</sup>.

# 1.2.4 NO<sub>2</sub> Conversion Model

When atmospheric hydrocarbons and NO<sub>x</sub> is exposed to the sunlight for a prolonged period, it undergoes complex photo chemical reactions, including oxidation of NO, decomposition of NO<sub>2</sub> and creation of oxidants. When the diffusion time is limited to a short period of one to two hours, however, nitrogen oxides (mostly NO) that are emitted from a stack or car will react with ozone as

they diffuse to form  $NO_2$ . The main reactions involved here are oxidation of NO by ozone and dissociation into NO by ultraviolet rays. Fig. 1-19 shows the  $NO_x$  and  $O_3$  distributions in a smoke stream measured by a group of Washington University. <sup>15)</sup> It is seen that  $O_3$  concentration was low where  $NO_x$  concentration was high, indicating the reaction of NO with  $O_3$ .

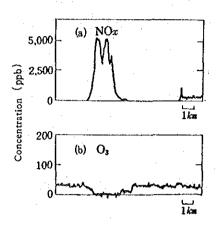


Fig. 1-19 NO<sub>x</sub> and O<sub>3</sub> Concentration in Plume Concentration (ppb)

For use in combination with the plume or puff diffusion model to calculate NO and NO<sub>2</sub> concentrations, Yamamoto and Yokoyama<sup>16)</sup> developed the following simple empirical formula from field test results.

$$C_{NO_2} = C_{NO_X} \{1 - \alpha e^{-kt}\}$$
 (1-21)

Here,  $C_{NO_x}$  is the NO<sub>x</sub> concentration obtained from the diffusion calculation equation and the concentration is assumed to be maintained.  $\alpha$  and k are test constants. In Equation (1-21),  $C_{NO_x}/NO_x$  becomes 100% when  $t\rightarrow\infty$ . In the nighttime, however, an equilibrium is attained with  $\frac{[NO] \cdot [O_3]}{[NO_2]} = Const.$ ) due to oxidation by ozone, dissociation into NO by ultraviolet rays, etc. and 100% conversion of NO into NO<sub>2</sub> does not take place. In the nighttime, 100% conversion is considered possible if ozone is present because of an absence of dissociation into NO. In the actual atmosphere, however, the  $NO_2/NO_x$  ratio in the nighttime remains nearly the same at 80% or so as during the daytime. This may be explained as follows: during nighttime, no ozone is produced but it is supplied from the higher atmosphere. When ozone so supplied is consumed in the reaction with  $NO_x$ , no further reaction takes place. The fact that both ozone and NO are found in the measurement of the atmosphere would be because bulk of pollutants exists heterogeneously both in terms of space and time and the average calculated reflects them. Therefore, in this study, Equation (1-22) was used to calculate  $NO_2$  concentration instead of Equation (1-21).

$$C_{NO_2} = C_{NO_2} \left\{ 1 - \frac{\alpha}{1+\beta} (e^{-kt} + \beta) \right\}$$
 (1-22)

where:

C<sub>NO</sub>: NO<sub>x</sub> concentration calculated by the diffusion calculation equation

 $\alpha$ : NO/NO<sub>x</sub> ratio in initial stage of diffusion ( $\alpha$ =0.9)

 $\beta$ : Constant to suppress NO to  $NO_2$  conversion ( $\beta$ =0.3)

k: Apparent reaction speed constant

k=0.0062·u·Co<sub>3</sub> (Stationary source)

k=0.062·u·Co<sub>3</sub> (Stationary area source, ship)

 $k=0.208 \cdot u \cdot Co_3$  (Car)

Co3: Ozone background concentration

u: Wind velocity (m/s)

As for the ozone background concentration, the values (one-hour values) measured at the roof of ONEB for a period of one year in 1988 were grouped by the atmospheric stability, and their medians were used. When calculating  $NO_2$  atmospheric diffusion concentration from the road, the ozone background concentration was corrected by Equation (1-23)<sup>10)</sup>. This is to take account of the decrease of ozone supplied to the background when the wind direction is parallel to the road (when the angle between the road and wind direction is 30 degrees or smaller).

Table 1-13 Ozone Background Concentration

Ozone	1	time .	Nighttime			
concentration	Unstable	Neutral	Neutral	Unstable		
(ppm)	0.014	0.010	0.005	0.005		

Daytime 
$$[O_3]_{BG}$$
 parallel wind= $0.55 \times [O_3]_{BG}$   
Nighttime  $[O_3]_{BG}$  parallel wind= $0.33 \times [O_3]_{BG}$  (1-23)

#### 1.2.5 Annual Average Value Model

The diffusion calculation was conducted for each categorized meteorological condition (combination of wind direction, wind velocity and atmospheric stability) and, from the calculated results and the frequency of occurrence of respective meteorological conditions, the annual average concentration was calculated as shown below.

(1) Calculation of Average Concentration under the Windy Condition

$$\overline{C} = \sum_{i} \sum_{j} \sum_{k} C(D_i, V_j, S_{tk}) \cdot f(D_i, V_j, S_{tk}) \dots$$
(1-24)

where:

 $C(D_i, V_j, S_{tk})$ : One-hour concentration with wind direction  $D_i$ , wind velocity  $V_j$ , stability class  $S_{tk}$ 

 $f(D_i, V_j, S_{tk})$ : Frequency of occurrence with wind direction  $D_i$ , wind velocity  $V_j$ , stability class  $S_{tk}$ 

(2) Calculation of Average Concentration under the Calm Condition

$$\overline{C} = \sum_{k} C(S_{tk}) \cdot f(S_{tk}) \qquad (1-25)$$

where;

 $C(S_{tk})$ : Concentration with stability class  $S_{tk}$ 

 $f(S_{tk})$ : Frequency of occurrence with stability class  $S_{tk}$ 

These calculations under the windy and calm conditions were made for each season and time period and the average was determined taking the frequency of occurrence into consideration.

## 1.2.6 Comparative Study on Calculated and Measured Values

Calculation of ambient pollutant concentrations were done by using all SO<sub>2</sub> and NO<sub>x</sub> data from sources (factories, cars, vessels and ferries) opereating in Samut Prakarn province in 1988 and agreement between the measured values and the calculated values at the monitoring station was checked. This calculation was done five times in total until a good agreement was obtained as shown in Table 1-14.

#### (1) Evaluation of Agreement on Calculated and Measured Values

The yearround values of measured and calculated concentrations by season and by time zone are listed in Table 1-15. The distribution of the measured and calculated values is shown in Fig. 1-20. The result of evaluation of these calculated values in terms of the criteria<sup>17)</sup> of the applied simulation model shown below is given in Table 1-16. From these results, it was found that the agreement of the yearly concentrations drops in the rank A except that of SO<sub>2</sub> (during the night-time zone) which is the rank C, and thus the simulation model used in this study is considered having a feasible accuracy.

① The difference  $(a_0)$  between the average  $(\overline{Y})$  of measured values and the average  $(\overline{X})$  of calculated values should satisfy the following condition.

$$\mathbf{a}_0 \le \left\{ \frac{1}{3} (\overline{\mathbf{Y}} - \mathbf{B}\mathbf{G}) + \mathbf{B}\mathbf{G} \right\}$$

BG: Background concentration

② The  $a_0$  shall be within the following range.

$$\mathbf{a}_0 \le \left\{ \frac{2}{5} (\mathbf{Y} - \mathbf{B}\mathbf{G}) + \mathbf{B}\mathbf{G} \right\}$$

- (3) The correlation coefficient must be more than 0.71 and exceed 0.80 (preferably bigger). The regression coefficient (a) must be from 0.8 to 1.2.
- 4) The estimated error rate  $(S'/\overline{Y})$  must not exceed 1/5.

- (5) S'/Y must not exceed 1/4.
- (6)  $S'/\overline{Y}$  must not exceed 1/3.

The ranks are established by combining criteria (1) to (6) as shown in the following table.

Evaluation Rank	Criteria
A	①, ③ and ⑤, or ① and ④ satisfied.
В	② and ⑤ satisfied.
C	② and ⑥ satisfied.

The following evaluation ranks are desirable, with respect to season and time period.

		Seasonal	Category
to the second second		Each Season	Whole Year
Time Zone Category	Each Time zone	C	В
	All day	В	Α

Table 1-14 Process of SO<sub>2</sub>, NO<sub>x</sub> and NO<sub>2</sub> Environmental Concentration Simulation Improvement

<u></u>	Condition	First	Second	Third Fourth	Fifth			
	Diffusion width	P-G diagram	HIWA	Y diagram	The diffusion width was deviated a half rank to the unstable size from the HIWAY diagram			
	Weight of diesel car	Heavy 5.5 t Light 1,62 t	Heavy 9.75 t Light 1.5 t					
1	σ <sub>20</sub> (Car)	3 m	5 m					
	He (Car)	2 m	3 m					
	Ratio of upper layer and surface wind velocity	EPA value (10 m→2 m)	EPA value (10 m→3 m)	Sam	e as left			
ž	Reaction coefficient	$a=0.80$ $\beta$ daytime=0.3 $\beta$ nighttime=0.0 $k=0.23$	α=0.9 β=0.3 k=0.208					
	Ozone reduction under parallel wind direction	Not considered	Considered	· ·				
	S content of diesel oil	0.569	% (Average of 0.46% and 0.6	5%)	0.46%			
	Light freight-train car			Diesel and gasoline cars 3:1				
	Transportation volume in the vicinity around MS5	Value measured in t	Modified by the ratio of two points on the Bangna-Trat highway to 15,577 vehicles/ day					
	Diffusion width	P-G diagram	Yamamoto/Yokoyama type					
	Effective stack height under windy condition	Large-scale stack; Moses & Carson type Small stack; CONCAWE type	CONCAWE type					
	Reaction coefficient	$\alpha$ =0.8 $\beta$ daytime=0.3, $\beta$ nighttime=0.0 k; Point sources=0.0062 k; Area sources=0.062 k; Vessel=0.062	α=0.9 β=0.3 k is same as left.	Samo	e as left			
Factory/vessel	Down wash	Not considered	Considered (Stack with top shade is always down-washed.)	Not co	: onsidered			
Facto	He of vessel/ferry	Vessel: 20 m; Ferry: 5 m	Same as left	Vessel; 30 m Ferry anchorage; 10 m Ferry navigation; 15 m				
	Area sources of He		30 m	40 m	30 m			
	Parameter k of σ <sub>E</sub>	A, B C 2.7 1.0	D E F 1.0 0.83 0.83		C D E F 1.04 1.0 0.83 0.83			
	S content of fuel used with vessel	2.5%; unc	Less than 1,000 tonnage: 1,03% 1,000 Tonnage or greater: 2,5%					
	S content of low speed diesel oil used with ferry boat	2.5%; under the	assumption that low speed d	iesel oil is used	0.56%: under the assumption that high speed diesel oil is used			
	Ferry boat horsepower	240	PS for F1, F2 and F3 course	es	F1, F2; 225 PS F3 ; 200 PS			

Table 1-15 Comparison of Measured and Calculated Values of Ambient SO<sub>2</sub>, NO<sub>2</sub> and NO<sub>x</sub> Concentration by Season and Time Zone

(SO<sub>2</sub>)

(unit;ppb)

STATION	So	uth We	t	No	rth Dr	У	So	uth Dr	у	A	nnua l	
2191100	Day	Nig.	Ave									
O(C1) OURD COLOTION	6.4	7.1	6.8	5.9	11.8	9.1	4.6	5.3	4.9	5.8	7.8	6.9
(MS1) ONEB STATION	7.1	12.8	10.2	4.3	9.0	6.8	5.7	8.3	7.1	6.1	10.7	8.6
(MS2) POWER PLANT	9.3	12.9	11.3	8.5	19.7	14.7	9.7	12.8	11.4	9.2	14.6	12,2
(IDZ) FUWER FLAMI	10.2	12.0	11.2	9.6	16.5	13.4	9.2	7.2	8.2	9.8	12.0	11.0
(MS3) MIN.DEP.OFFICE	17.1	23.5	20.6	15.9	22.6	19.6	20.9	43.3	33.1	17.7	28.3	23.5
(FLOG) MIN. DEF. OFFICE	20.8	25.1	23.1	16.4	16.8	16.6	20.5	16.3	18.2	19.6	20.8	20.3
ONCY) & D DDO OFFICE	3.6	5.1	4,4	5.3	8.7	7.1	3.8	4.1	4.0	4.0	5.7	5.0
(MS4) S.P.PRO.OFFICE	6.5	9.5	8.1	6.5	12.1	9.6	5.5	5.9	5.7	6.3	9.3	7.9
AICE) II O T ECTATE	3.5	4.4	4.0	1.8	3.0	2.5	1.6	1.7	1.7	2.4	3.2	2.9
(MS5) H.& I. ESTATE	1.5	2.2	1.9	2.2	2.9	2.6	1.0	1.5	1.3	1.6	2.2	1.9

(NO₂

STATION	So	uth We	t	No	rth Dr	у	So	uth Dr	y Annual			
SIRITON	Day	Nig.	Ave	Day	Nig.	Ave	Day	Nig.	Ave	Day	Nig.	Ave
(MS1) ONEB STATION	17.8	17.0	17.4	10.9	21.1	16.5	10.0	11.7	11.0	14.1	16.7	15.6
(DI) ONED DIATION	10.9	14.8	13.0	11.8	18.4	15.4	8.4	10.4	9.5	10.5	14.6	12.7
(MS2) POWER PLANT	7.3	8.1	7.7	7.2	14.3	11.1	7.6	9.5	8.6	7.3	9.8	8.7
	5.8	9.0	7.5	4.0	8.7	6.5	5.4	6.3	5.9	5.3	8.2	6.9
(MS3) MIN.DEP.OFFICE	6.9	9.7	8.4	14.6	24.8	20.2	10.4	16.1	13.5	10.1	15.7	13.2
(455) MIN.DEL.OFFICE	9.6	9.7	9.7	8.1	7.7	7.9	9.3	6.9	8.0	9.2	8.5	8.8
(MS4) S.P.PRO.OFFICE	10.3	11.1	10.7	27.1	34.7	31.2	9.2	10.0	9.7	13.5	15.7	14.7
antrio.oni.i.c (Par)	10.0	11.3	10.7	11.8	18.0	15.2	8.3	7.4	7.8	10.0	12.0	11.1
(MS5) H.& I. ESTATE	3.4	5.7	4.7	4.1	8.3	6.4	3.5	5.2	4.4	3.6	6.1	5.0
(DJ) II.G I. ESIME	2.6	4.2	3.5	4.2	6.4	5.4	1.9	3.5	2.8	2.8	4.6	3.8

(NO<sub>x</sub>)

<u></u>												
STATION	So	uth We	t	No	rth Dr	у	So	uth Dr	y	A	nnua l	
DIMITON	Day	Nig.	Ave									
(MS1) ONEB STATION	39.7	56.4	48.8	20.3	47.7	35.3	17.2	19.9	18.7	29.2	45.1	37.9
(MDI) UNED STATION	25.0	43.4	35.0	24.2	54.4	40.6	16.8	25.2	21.4	22.8	41.7	33.0
(MS2) POWER PLANT	16.2	16.7	16.5	15.0	24.4	20.1	17.0	18.0	17.6	16.2	18.7	17.6
(FDZ) TOWER TIZMI	11.8	15.5	13.8	7.1	14.1	10.9	11.6	10.3	10.9	10.6	13.9	12.4
(MS3) MIN.DEP.OFFICE	14.3	20.0	17.4	25.5	44.4	35.8	19.1	26.0	22.8	18.8	28.5	24.1
appriorize (ecr)	23.9	24.9	24.5	21.3	21.4	21.3	22.9	17.7	20.1	23.0	22.3	22.6
(MS4) S.P.PRO.OFFICE	24.4	28.6	26.7	49.8	71.7	61.7	25.5	25.1	25.3	29.9	36.7	33.6
3.1.TNU,UNTICE	24.3	30.9	27.9	27.3	47.0	38.0	21.2	21.6	21.4	24.3	32.7	28.8
(MS5) II.& I. ESTATE	8.3	12.0	10.3	6.8	11.5	9.4	5.5	7.1	6.3	7.2	10.5	9.0
GDO) HIG I. ESIMIE	5.3	9.1	7.4	9.7	14.0	12.0	4.0	7.2	5.8	6.1	9.9	8.1

UPPER ; observed concentration LOWER ; calculated concentration

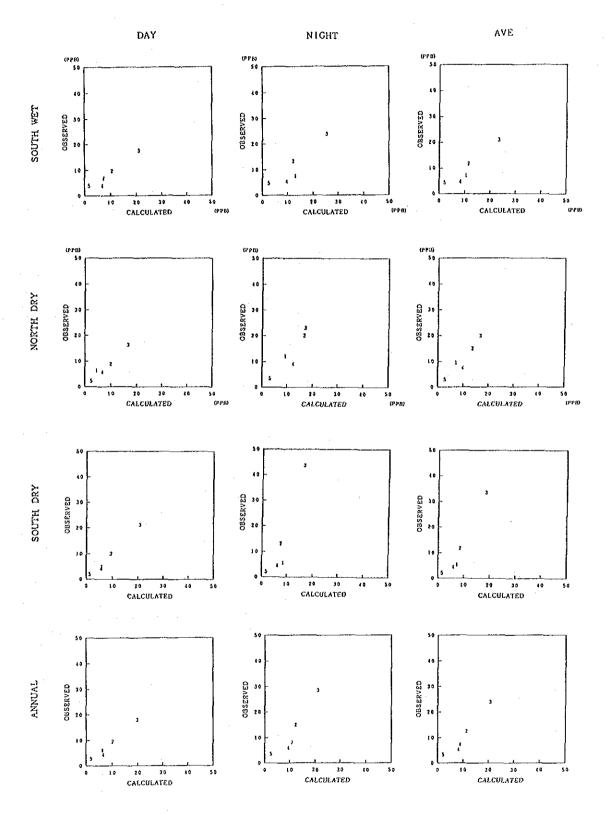


Fig. 1-20(1) Comparison of Measured and Calculated Values of SO<sub>2</sub> Ambient Concentration

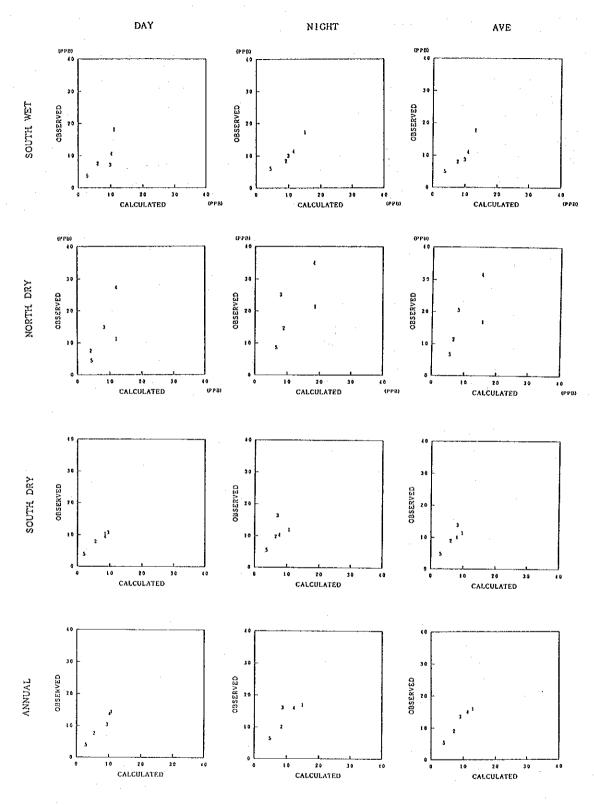


Fig. 1-20(2) Comparison of Measured and Calculated Values of NO<sub>2</sub> Ambient Concentration

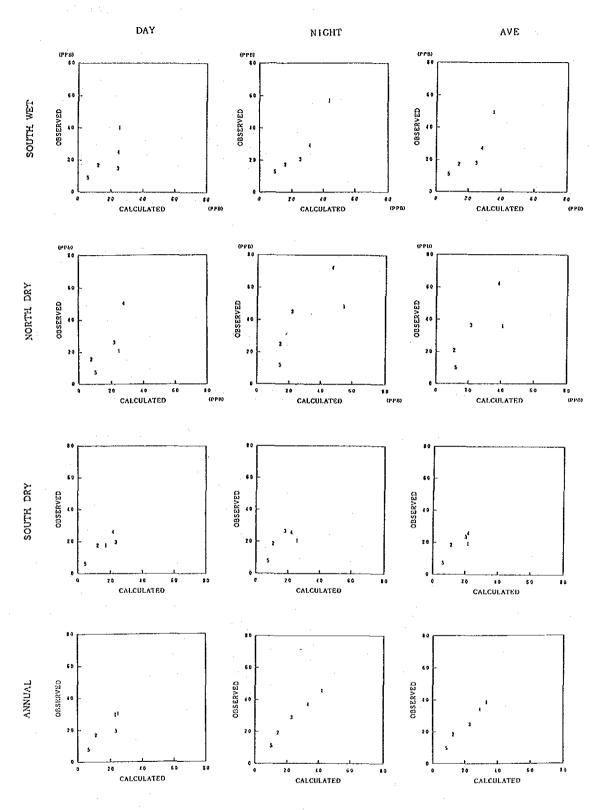


Fig. 1-20(3) Comparison of Measured and Calculated Values of  $NO_{\rm x}$  Ambient Concentration

Table 1-16 Evaluation of Agreement of Ambient Pollutant Concentration Simulation Model

# [802]

Season an time		Ave.of obs. Y (ppb)	Ave.of cal. X (ppb)	A <sub>0</sub> (Y-X) (ppb)	Regression equation	Corr.	Vari. coef.	Rank
	Day	7.9	9.2	-1.3	Y = 0.760X + 1.0	0.969	0.25	В
South Wet	Night	10.6	12.3	-1.7	Y = 0.881X - 0.3	0.914	0.29	С
	Ave.	9.4	10.9	-1.5	Y = 0.846X + 0.2	0.947	0.24	Α
	Day	7.5	7.8	-0.3	Y = 0.936X + 0.2	0.978	0.14	Α
North Dry	Night	13.2	11.5	1.7	Y = 1.272X - 1.4	0.922	0.24	В
	Ave.	10.6	9.8	0.8	Y = 1.162X - 0.8	0.957	0.18	Α
	Day	8.1	8.4	-0.3	Y = 1.039X - 0.6	0.991	0.12	Α
South Dry	Night	13.4	_7.8	5.6	Y = 2.949X - 9.7	0.922	0.83	
.,	Ave.	11.0	8.1	2.9	Y = 1.994X - 5.1	0.966	0.57	
	Day	7.8	8.7	-0.8	Y = 0.885X + 0.2	0.987	0.14	Α
Annual	Night	11.9	11.0	0.9	Y = 1.412X - 3.6	0.936	0.33	C
	Ave.	10.1	9.9	0.1	Y = 1.197X - 1.8	0.968	0.22	Α

# [NO<sub>2</sub>]

Season an time		Ave.of obs. Y (ppb)	Ave.of cal. X (ppb)	A <sub>0</sub> (Y-X) (ppb)	Regression equation	Corr.	Vari. coef.	Rank
: 1	Day	9.1	7.8	1.4	Y = 1.203X - 0.2	0.776	0.34	_
South Wet	Night	10.3	9.8	0.5	Y = 1.052X + 0.0	0.953	0.11	Α
	Ave.	9.8	8.9	0.9	Y = 1.197X - 0.8	0.908	0.19	Α
	Day	12.8	8.0	4.8	Y = 1.741X - 1.1	0.750	0.46	: —
North Dry	Night	20.6	11.8	8.8	Y = 1.155X + 7.0	0.677	0.32	С
	Ave.	17.1	10.1	7.0	Y = 1.426X + 2.7	0.729	0.36	_
	Day	8.1	6.7	1.5	Y = 0.912X + 2.1	0.989	0.05	Α
South Dry	Night	10.5	6.9	3.6	Y = 0.946X + 4.0	0.586	0.27	С
	Ave.	9.4	6.8	2.6	Y = 1.131X + 1.8	0.872	0.16	Α
	Day	9.7	7.6	2.2	Y = 1.272X + 0.1	0.974	0.13	Α
Annua 1	Night	12.8	9.6	3.2	Y = 1.032X + 2.9	0.856	0.17	A
	Ave.	11.4	8.7	2.8	Y = 1.238X + 0.7	0.976	0.10	Α

# [NOx ]

Season an time		Ave.of obs. Y (ppb)	Ave.of cal. X (ppb)	A <sub>0</sub> (Y-X) (ppb)	Regression equation	Corr.	Vari. coef.	Rank
	Day	20.6	18.1	2.5	Y = 0.942X + 3.6	0.696	0.38	. —
South Wet	Night	26.7	24.8	2.0	Y = 1.238X - 3.9	0.939	0.23	В
	Ave.	23.9	21.7	2.3	Y = 1.192X - 1.9	0.874	0.28	C
	Day	23.5	17.9	5.5	Y = 1.432X - 2.2	0.793	0.41	
North Dry	Night	39.9	30.2	9.8	Y = 0.948X + 11.3	0.786	0.32	С
	Ave.	32.5	24.6	7.9	Y = 1.119X + 5.0	0.797	0.33	С
1	Day	16.8	15.3	1.5	Y = 0.830X + 4.1	0.883	0.19	A
South Dry	Night	19.2	16.4	2.8	Y = 0.719X + 7.4	0.716	0.27	С
	Ave,	18.1	15.9	2.2	Y = 0.896X + 3.9	0.881	0.17	Α
	Day	20.2	17.4	2.9	Y = 0.999X + 2.9	0.832	0.20	Α
Annual	Night	27.9	24.1	3.8	Y = 1.035X + 3.0	0.989	0.07	Α
	Ave.	24.4	21.0	3.5	Y = 1.094X + 1.5	0.988	0.08	A

## (2) Contribution Rate by Source at Monitoring Station

The measured values at each station are as shown in Table 1-15. The contribution rate by source (factory, car, vessel and ferry) was calculated while accounting for the background concentration.

The difference between the average y of the measured concentration and the average x of the calculated concentration at each monitoring station is as shown in Table 1-16. The difference is thought due to the following causes.

- ① Pollutants emitted from sources in the target area or in adjacent area are made blown back by circulating sea or land breeze or due to change of wind direction.
- ② Pollutants emitted from middle/small sources or moving sources which cannot be detected in the target area or in adjacent area.
- (3) Pollutants long-term staying in the ambient air
- (4) Measurement error at monitoring station
- (5) Adsorption of pollutants to ground surface
- 6 Decomposition of pollutants in atmosphere
- (7) Background concentration from natural sources

In this study, this  $\overline{y}-\overline{x}$  was added to the calculated value as background concentration. Table 1-17 shows the contributing concentrations and their contribution rate by source to yearly average concentration of each monitoring station while the background concentration is taken into consideration.

The significant contribution of the sources to SO<sub>2</sub> concentration were found in small factories having the stacks less than 50 m high and in automobiles. The sizable plants with stacks more than 50 m high have small contribution less than 10 pct at most. As for NO<sub>2</sub>, the largest contributer is automobiles and next comes ships. The factories large and small combined have contribution of 10 pct or so.

Table 1-17 Contribution Concentration by Source at Monitoring Station

80:

	Obser~				yea	rly	average	conc	entratio	n an	d ratio	for	each sou	гсе	· :			
Station	ved conc.	Total	Station (point H• ≧8		Station (point H. <5	агу Оз	Station ( area		Roadwa	ys	Vessel saili		Ferrybo anchor		Ferrybo saili		Back grou	nuq
	(ppb)	(ppb)	(ppb)	%	(ppb)	%	(ppb)	%	(ppb)	%	(ppb)	%	(ppb)	%	(ppb)	%	(ppb)	%
OVS1) OVER STATION	6.9	8.7	0.25	3	2.27	26	2.08	24	3.76	43	0.25	3	0.00	0	0.00	0	0.1	1
(MS2) POWER PLANT	12.2	11.1	1.01	9	3.97	36	3.68	33	0.95	9	1.35	12	0.00	0	0.01	0	0.1	1
OS3) NIN. DEP. OFFICE	23.5	20.4	0.47	2	10.14	50	4.26	21	0.85	4	3.22	16	0.10	1	1.22	6	0.1	0
(MS4) S.P.PRO.COFICE	5.0	8.0	0.07	1	2.00	25	1.55	19	3.46	43	0.82	10	0.00	0	0.00	0	0.1	1
OISS) H. & I. ESTATE	2.9	2.0	0.01	0	0.32	16	0.42	21	1.14	57	0.01	0	0.00	0	0.00	0	0.1	5

NO<sub>2</sub>

	l				yea	rly	average	conc	entratio	n an	d ratio	for	each so	rce				
Station	Obser- ved conc.	Total	Station (point He ≧5		Station (point No <5	ary Oa	Station (area		Roadwa	ys	Vessel saili		Ferrylo anchor		Ferrybo saili		Back grou	indi
	(ppb)	(ppb).	(ppb)	%	(pob)	%	(ppb)	%	(ppb)	%	(pob)	%	(ppb)	%	(ppb)	%	(ppb)	%
(MS1) ONEB STATION	15.6	15.5	0.18	1	0.09	1	0.17	1	12.00	77	0.29	2	0.00	0	0.01	0	2.8	18
(MS2) POWER PLANT	8.7	9.7	0.20	2	0.13	1	0.35	4	5.27	54	0.86	9	0.01	0	0.05	1	2.8	29
(MS3) MIN.DEP.OFFICE	13.2	11.6	0.30	3	0.25	2	0.57	5	4.08	35	1.72	15	0.15	1	1.74	15	2.8	24
(NS4) S.P.PRO.OFFICE	14.7	13.9	0.04	0	0.09	. 1	0.12	1	10.25	74	0.60	4	0.00	0	0.01	0	2.8	20
(XS5) H. & I. ESTATE	5.0	6.6	0.01	0	0.02	0	0.02	O	3.71	57	0.01	0	0.00	0	0.00	0	2.8	43

NO<sub>x</sub>

NOx					100		:						100		200			2.5
			-		yea	rly	average	conc	entratio	ns nx	d ratio	for	each sou	rce	:			
	Obser- ved	7-1-1			Station	άŢΥ	Station	згу			Vessel	S	Ferrybo	als	Ferrybo	als	Back	
Station	conc.	Total	(point	0á	point  > e	Oa	( area	)	Roadw	ıys	saili	ng	anchos	ing	saili	ng	grou	ınd
	(699)	(ppb)	(ppb)	%	(ppb)	%	(ppb)	%	(ppb)	%	(ppb)	%	(ppb)	%	(ppb)	%	(cpb)	%
OIST) ONER STATION	37.9	36.5	0.63	2	0.41	.1	0.30	1	31.22	85	0.44	1	0.00	0	0.01	0	3.5	10
(MS2) POWER PLANT	17.6	15.9	1.42	9	0.66	4	0.72	5	7.06	45	2.42	15	0.01	0	0.07	.0	3.5	22
(MS3) MIN.DEP.OFFICE	24.1	26.1	1.24	5	1.55	6	1.14	4	6.12	21	5.77	22	0.46	2	6.33	24	3.5	13
(%S4) S.P.PRO.OFFICE	33.6	32.3	0.18	l	0.42	1	0.22	ı	26.54	82	1.46	4	0.00	0	0.01	0	3.5	11
(MSS) II, & I. ESTATE	9.0	11.6	0.02	0	0.03	1	0.05	0	7.97	69	0.01	0	0.00	0	0.00	0	3.5	30

# 1.3 Estimation of Ambient Pollutant Concentrations for Whole Area and Contribution Rate by Source by Using Atmospheric Diffusion Simulation Model

With the atmospheric pollutant sources and meteorological condition of 1988 as input data,  $SO_2$ ,  $NO_2$  and  $NO_x$  environmental concentrations were calculated by season, by time zone and by yearly period for the whole Samut Prakarn province (1 km×km mesh). As for the  $SO_2$  and  $NO_2$  yearly average values, the contribution rate by source to points with significant concentration level was calculated.

# 1.3.1 SO<sub>2</sub>, NO<sub>2</sub> and NO<sub>x</sub> Concentrations in whole Samut Prakarn Province

Fig. 1-21 shows the yearly average concentrations by season and by time zone in the whole Samut Prakarn province in 1988 expressed as a topograhic diagram. This calculated value includes background concentration.

# 1.3.2 Comparison with Environmental Standards

The yearly average concentrations in the whole area and the environmental standard were compared. For the environmental standard in Thailand, SO<sub>2</sub> is determined by daily average value and geometric yearly average value, and NO<sub>2</sub> is determined by one-hour average value. However, the calculated value of ambient pollutant concentration is the yearly average value, and so it is necessary for calculated value and the environmental standard value to be compared on the same conditions. In this study, the relationship of the yearly average of the measured values in each monitoring station, the daily average value of SO<sub>2</sub> and one-hour value of NO<sub>2</sub> were expressed by linear regression equation of the first order. Then, the yearly average value calculated from this line was converted to the equivalent daily average or one-hour average value, which was then compared with the environmental standards.

Fig. 1-22 shows the relationship between the yearly and max. of daily average values of  $SO_2$  and the relationship between the yearly and max. of one-hour average values of  $NO_2$ . In this figure, if the yearly average value of  $SO_2$  exceeds 42 ppb, or if the yearly average value of  $NO_2$  exceeds 18 ppb, the environmental standards will be violate.

Fig. 1-23 shows the calculation results of SO<sub>2</sub>, NO<sub>2</sub> and NO<sub>x</sub> yearly averages. Table 1-18 shows number of meshes by rank of concentration. From the table, both SO<sub>2</sub> and NO<sub>2</sub> concentrations are found satisfying the environmental standards across the whole Samut Prakarn Province.

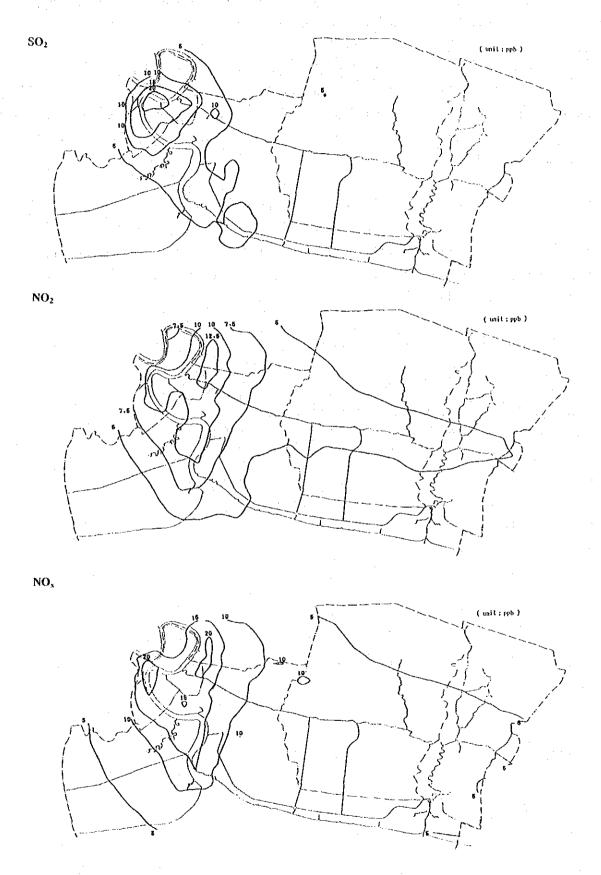


Fig. 1-21 Yearly Average Ambient Pollutant Concentrations in the Samut Prakarn Province in 1988

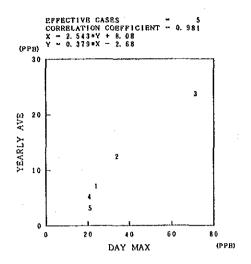


Fig. 1-22(1) Relationship between SO<sub>2</sub> Yearly Average Concentration and Max. of Daily Average Concentration

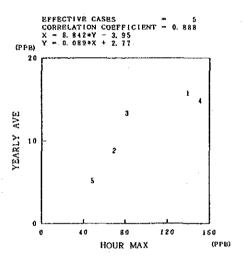


Fig. 1-22(2) Relationship between NO<sub>2</sub> Yearly Average Concentration and Max. of One-hour Average Concentration

Table 1-18 Number of Meshes by Rank of Concentration

(802)		(NO <sub>2</sub> )	1	(NOx)	)	
Rank of cocentration (ppb)	Number of mesh	Rank of cocentration (ppb)	Number of mesh	Rank of cocentration (ppb)	Number of nesh	
0.0 ~ 5.0	977	0.0 ~ 5.0	674	$0.0 \sim 5.0$	277	
5.0 ~ 10.0	128	5.0 ~ 7.5	305	5.0 ~ 10.0	676	
10.0 ~ 15.0	28	7.5 ~ 10.0	109	10.0 ~ 15.0	116	
15.0 ~ 20.0	21	10.0 ~ 12.5	64	15.0 ~ 20.0	79	
20.0 ~ 25.0	5	12.5 ~ 15.0	7	20.0 ~ 25.0	9	
TOTAL	1159	TOTAL	1159	25.0 ~ 30.0	2	
		<u></u>		TOTAL	1159	

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Fig. 1-23(2) NO<sub>2</sub> Concentration Distribution in Samut Prakarn Province in 1988

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Fig. 1-23(3) NO<sub>x</sub> Concentration Distribution in Samut Prakarn Province in 1988

# 1.3.3 Estimation of Contribution Rate by Source

Because both SO<sub>2</sub> and NO<sub>2</sub> concentrations are found to meet the environmental standards at Samut Prakarn, top eight points that have high concentration are selected and the contribution rates of SO<sub>2</sub> and NO<sub>2</sub> yearly average concentrations by source at these points are calculated. The eight selected points are as shown in Table 1-19 and in Fig. 1-24.

Table 1-19 High Cocentration Point and Yearly Average Concentrations of SO<sub>2</sub> and NO<sub>2</sub>

		(802)
SEQ)	Hesh Hesh	yearly average concentration (pob)
<del> </del>		
1)	16 - 24	22, 9
2)	16 - 23	22. 5
3)	17 - 23	21.9
4)	15 - 23	20, 8
5)	17 - 24	20.4
6)	15 - 22	19. 7
7)	18 - 23	1.9. 6
8)	15 - 21	19.0

		(NO <sub>2</sub> )
SEQ)	Kesh Kesh IX IY	yearly average concentration (ppb)
1)	22-26	14, 3
2)	22-25	13.8
3)	22 - 27	13.7
4)	21 - 24	13, 6
5)	21 - 23	13.4
G)	22 - 24	13.3
7)	21 - 25	12, 8
8)	15-24	12.4

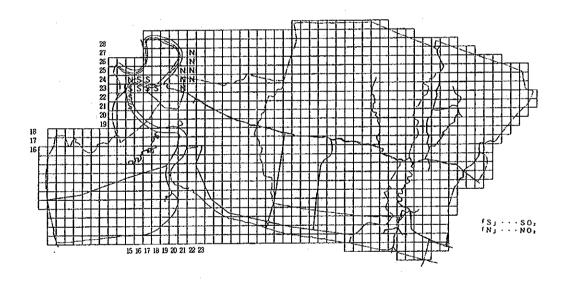


Fig. 1-24 High Concentration Points Applied for Calculation of Contribution Rate

Table 1-20 shows the contribution concentration and the contribution rate by source at the high concentration points of  $SO_2$  and  $NO_2$ . In this table, 10 sources with higher contribution concentrations are listed. The results are summarized as follows.

The high annual average concentrations of SO<sub>2</sub> and NO<sub>2</sub> appear in the vicinity of MS3 and MS1 respectively. The contribution of the sources are in case of SO<sub>2</sub>, factories (81.7–88.1 pct), auto-

mobiles (3.8-5.6 pct), ships (6.3-11.8 pct), ferries (0.1-8.0 pct). As for  $NO_2$ , they are factories (3.0-8.4 pct), automobiles (28.6-74.6 pct), ships (2.4-12.2 pct), and ferries (0.1-28.2 pct). In other words, the largest contributer of  $SO_2$  and  $NO_2$  are factories and automobiles respectively.

When the sources which are accountable for SO<sub>2</sub> and NO<sub>2</sub> are examined from the point of stack height, small groups of stationary sources which are about 10 meters high account for the largest percentage. Other than United Grain Co., Ltd. (3-75-1) in regard to SO<sub>2</sub>, there is no single source which is accountable for more than 10 percent.

Table 1-20(1) Contribution Concentration (Yearly Average Value) at High Concentration Points of SO<sub>2</sub>

2)	Conce. Rate		90000000000000000000000000000000000000	5.05 25.8	5.433 2.2513 6.52513 6.5254 6.132 6.	19.684 100.0	21)	Conce. Rate		20000000000000000000000000000000000000	§.83. 28.3	0.000 0.000	18.986 100.0
( Rank 6 IX, IY=15,22)	Type of source	SBB) Coun Fact Sta. c	5.7. 9.5. 	Remaining stacks Sub total	Factory (area source) Roadways Vessels (sailing) Ferryboats (anchoring) Ferryboats (sailing) Back ground		[ Rank 8 IX, IY=15,21]	Type of source	SEQ) Coun Fact Sta. (		Remaining stacks Sub total	Factory (area source) Readways Vessels (salling) Ferryboats (anchoring) Ferryboats (sailing) Back ground	
	Rate		03000000000000000000000000000000000000	2583 656	240000 640000 6600000000	1 100.0			) (%):-	######################################	8 25.0 1 61.8	7-05:00 64:00:00 7-00:00-10	8 100.0
Z,	Sice	contr. (pb)		14.8%	1000-1000 1000-1	20.441	ଛି	Sign	contr. (ppb)	-0000000000000000000000000000000000000	25.188	882823	19.618
( Rank 5 IX, IY=17, 24)	Type of source	SEQ) Coun Fact Sta.	-20040000000000000000000000000000000000	Regaining stacks Sub total	Pactory (area source) [oadways Vessels (sailing) Ferrybats (anchoring) Ferrybats (sailing) Back ground	TOTAL	( Rank 7 IX, IY=18, 23)	Type of source	SEQ) Coun Fact Sta.	### ##################################	Remaining stacks Sub total	Factory (area source) Roadways Vessels (sailing) Ferryboats (anchoring) Ferryboats (sailing) Back ground	TOTAL
•						_	1					·	· ·
	Rate	: <u>§</u>		25. 25.	%-400000 ಬೆಟ್ಟರು-ಸುವ	180.0		Rate	:#S	CONTROL CONTRA	22.03 62.00	ಬ್ಲಿಬ್ಲರೆ-4-0 ಬಹುದರ್ನು	0.00
ଛ	Sace.		20000000000000000000000000000000000000	35.583 5.583	00-000 80-000 000-000	22.496	ន	Conce.	Spirit Poblic		15.93	40%0-0 18668241 188682865	20.840
( Rank 2 IX, IY=16,23)	Type of source	) Coun Fact Sta.	503 900 503 900 503 600 600	Remaining stacks Sub total	Factory (area source) Roadways Vessels (salling) Ferryboats (ancipring) Ferryboats (salling) Back ground	TAL	( Rank 4 IX, IY=15,23)	Type of source	SEG) Coun Fact Sta.	できる。 - 1959 第5 - 1	Remaining stacks Sub total	Factory (area source) Roadways Vessels (sa)ling) Ferryboats (anchoring) Ferryboats (sailing) Back ground	TOTAL
	Rate	i. Si	ricking in	588 500	845000 ಬಂಬ-4-4	0.03		Rate	:: :::::::::::::::::::::::::::::::::::	addaminionio recocección	82.3	24-000 48-0-0	0.0
<b>3</b> 3	Conce.		40000000000000000000000000000000000000	4.788	0.000000 200000000000000000000000000000	22.886	(Si	<u>ور</u>	Society (pag)	20000000000000000000000000000000000000	13.152	######################################	21.832
( Rank 1 IX, IY=16, 24)	Type of source	SEQ) Coun Fact Sta.	50 75 50 75 7	Remaining stacks Sub total	Factory (area source) Readways Vessels (salling) Ferryboats (anchoring) Ferryboats (salling) Back ground	TOTAL	( Rank 3 IX, IY=17,23)	Type of source	SEQ) Coun Pact Sta.	5.5. 9.5. 5.5. 9. 5.5. 9. 5.5. 9. 5.5. 9. 5.5. 9.	Remaining stacks Sub total	Factory (area source) Roadways Vessels (salling) Ferryboats (anchoring) Ferryboats (salling) Back ground	TOTAL

Table 1-20(2) Contribution Concentration (Yearly Average Value) at High Concentration Points of NO<sub>2</sub>

	Rate	: ::	9000000000 40000	2.3	600000 10000000000000000000000000000000	180.0	:		Rate of	<del>S</del>	ಇದದ್ದಾದ್ದಾದ್ದರ್ ಒತ್ತಮಬಹುಬಹುಗ	2.4	485488 000000000000000000000000000000000	0.0
[X]	Conce.	contr. (ppb)	00000000000000000000000000000000000000	88 88	0000000 2500000000000000000000000000000	13.343		(%)	Conce.	contr (ppb)	255 355 355 355 355 355 355 355 355 355	0.581	00000000000000000000000000000000000000	12.445
( Rank 6 IX, IY=22, 24)	Type of source	SEQ) Coun Fact Sta.		Remaining stacks Sub total	Factory (area source) Roadways Vessels (sajling) Ferryboats (anchoring) Ferryboats (sailing) Back ground	TOTAL		( Rank 8 IX, IY=15,24)	Type of source	SEG) Coun Fact Sta.	\$\frac{\partial}{2}\$\parti	Remaining stacks Sub total	Factory (area source) Roadways Vessels (sajling) Ferryboats (anchoring) Ferryboats (sailing) Back ground	TOTAL
[	Pate	<u>ښ</u>	<u> </u>	0.8 2.8		100.0			Rate	Šį.	000000000	8.7	0.00-100-12 0.00-100-18	100.0
(S)	Conce.	Ç.	     	8.348	00000000000000000000000000000000000000	13.450		ন্ত্র	Conce.		25555555555555555555555555555555555555	8.83	20000000000000000000000000000000000000	12.834
( Rank 5 IX, IY=21,23)	Type of source	SEG) Coun Fact Sta.	\$5.0 9.5 \$5.0 9.5 \$5.0000000000000000000000000000000000	Remaining stacks Sub total	Factory (area source) Roadways Vessels (salling) Ferryboats (anchoring) Ferryboats (salling) Bock ground	TOTAL		( Rank 7 IX, IY=21, 25)	Type of source	SEQ) Coun Fact Sta.	できる。 - 2000年 - 20	Remaining stacks Sub total	Factory (area source) Roadways Vessels (sailing) Ferryboats (anchoring) Ferryboats (sailing) Back ground	TOTAL
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	Rate		0000000000 40000000000	2.1		100.0			Rate	1993	<u> </u>	8.8	2,-woog	100.0
ଥି	Conce	contr (ppp)		8:83	9500000 8128000000000000000000000000000000	13.808		প্র	Conce.	Contr. (ppb)		8:38	0.0000 800000 8000000 80000000000000000	13.600
( Rank 2 IX, IY=22	Type of source	SEQ) Coun Fact Sta.	5.5.6.5.5.6.5.5.6.5.5.6.5.5.6.5.5.6.5.6	Remaining stacks Sub total	Factory (area source) Roadways Vessels (salling) Ferryboals (anchoring) Ferryboats (salling) Back ground	TOTAL		( Rank 4 IX, IY=21,24)	Type of source	SEQ) Coun Fact Sta.	5,5 6.5 5,0 6.5 5,0 7.5 5,0 7.5 6,0	Remaining stacks Sub total	Factory (area source) Roadways Vessels (sailing) Ferryboats (anchoring) Ferryboats (sailing) Ferryboats (sailing) Back ground	TOTAL
	ate s	: :	00000000000000000000000000000000000000	2.0	1.5000.00 1.0000.00	100.0			Rate	9. 1.	90000000000000000000000000000000000000	2.0		100.0
( <u>%</u>	Conce.	contr. (Ppb)	99999999999999999999999999999999999999	8:88	9.55.05.9. 8.56.05.9. 8.56.05.9. 8.56.05.9.	14.272	·	52	Conce.		\$	8:38	2.0000000 8.0000000000000000000000000000	13.684
( Rank 1 1X,1Y=22,26)	Type of source	SEQ) Coun Fact Sta.	252 9 25 252 9 252 252 252 252 252 252 252 252 252 252	Remaining stacks Sub total	Factory (area source) Roadways Vessels (sailing) Ferryboats (anchoring) Ferryboats (sailing) Back ground	TOTAL	1	( Rank 3 IX, IY=22, 27)	Type of source	SEG) Coun Fact Sta.		Regaining stacks Sub total	Factory (area source) Radways Vessels (sailing) Ferryboats (anchoring) Rerryboats (sailing) Back ground	TOTAL
į	· 		Part Start	······································	<u> </u>			ļ			Factor (197)		~~ <u>~</u> ~~~	

# 2. Study on Atmospheric Pollutant Concentration of Particulate Matters Estimated by Chemical Mass Balance Method

# 2.1 Assumption Methods of Source Contribution Rate of Particulate Matter

To clarify the correlation between an emission source and the ambient concentration of the air pollutants, source models, such as a diffusion model, have been developed based on the condition that  $SO_2$  and  $NO_x$  emission volume from the source will be kept in the ambient in the same volume. However, as for the particulate matter, the same diffusion models cannot be applied because of the fact that of the particulate matter emitted from various sources, secondary particles are produced by chemical reaction in the ambient and modeling of the removal mechanism is not yet confirmed. It is generally understood that simulation methods are now under development. Among them, the receptor model has been developed which determines emission sources from various information of the ambient particulate matter such as concentration of the respective elements, ion concentration, particle size and forms.

In this chapter, the problem points of the source model applied for particulate matter and assumption methods of the source contribution rate by the receptor model are described.

# 2.1.1 Problem Points of the Source Model Applied for Particulate Matter

The source model is the method to estimate the ambient concentration and the contribution rate of each individual emission source (factory and stack) at certain points (mesh, station, etc.) based on the data on emission intensity, conditions (stack height, emission gas volume, etc.) and meteorological conditions (wind direction & velocity, atmospheric stability, etc.). For concentration calculation, the plume model, puff model and numerical model (diffusion factors are complicated) are widely used.

These diffusion models are applicable for environmental assessments which are necessary to establish environmental protection plans and location plans for new industries. These models can evaluate the above said plans and they can also evaluate the environmental impacts by the respective sources even when several sources are emitting the same types of pollutants. But as mentioned above, these models are not applicable for the assumption of ambient concentration and their contribution rate of the particulate matters due to the reasons mentioned hereafter.

#### (1) Ambient concentration

The instruments to measure short time (at least 1 hour) concentration of particulate matter by particle size are not yet developed. It is therefore impossible to clarify the correlation between meteorological data and particulate concentration.

#### (2) Emission source

It is difficult to clarify the wide range emission sources, and the generating mechanism of secondary particles quantitatively.

## (a) Stationary source

The weight concentration is easy to obtain, but the data on size distribution by facilities, type of fuels, operational conditions, dust collecting facilities, etc. are not easy to obtain.

### (b) Automobile

The particulate matter generated by automobiles consists of emission gases, the abrasion of tires, abrasion of pavement and particles blown up from the road. Therefore, it is very difficult to classify these particles quantitatively.

# (c) Ships and aeroplanes

It is difficult to identify not only the emission factor of weight concentration but also the size distribution.

# (d) Particulate generating facilities

It is difficult to obtain the data of emission volume from cement manufacturing, quarrying, steel mills, coal yards, reclamation sites, etc.

## (e) Natural background source

It is difficult to clarify the volume of particulate generated by volcano eruption, sea salt, dust storm, seeds of plants, etc. quantitatively.

#### (f) Others

Emission factors of outdoor burning of wastes, straw, etc. are unknown.

## (3) Prediction model

It is necessary to develop a new model which includes gravity fallout of the particulate, generation of secondary particles, removal mechanism of particulate, etc.

#### 2.1.2 Receptor Model

The receptor model is the method to identify the sources and their contribution rate based on the data on chemical composition, particle size, concentration variation, particle form, etc. which are obtained at a certain point and a certain period. This method has been recently highlighted due to the development of microanalysis techniques. Many studies on the receptor model have been carried out.

This method has the advantage of being able to identify the source and contribution rate without considering meteorological and geographical conditions, and also of being able to identify the new sources and sources for which data is difficult to obtain. Further, it is applicable for evaluation of emission sources which should be realized in case of high concentration, by using data measured. However, the receptor model is not effective for identification of each individual source when some emission sources emitting similar chemical substraits are existing as this model can be applied to one source type only, while the source model can be applied for evaluation of the ambient concentration that each individual emission source affects. The evaluation by receptor model is only limited to the area of monitoring stations and it is also impossible to predict the future conditions.

The receptor model can be classified to the morphological observation, physical analysis and chemical analysis, as shown in Fig. 2-1.

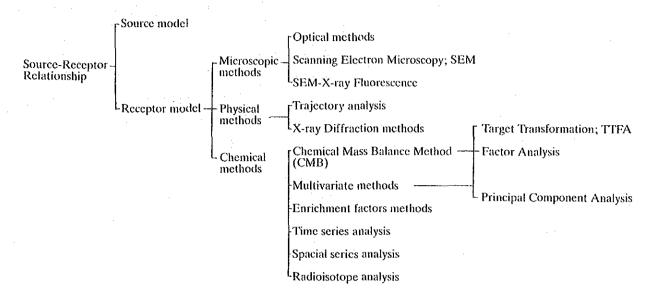


Fig. 2-1 Contribution Rate Assumption Methods of Particulate Matter

# (1) Morphological observation method

The morphological observation method assumes the source from the size of particulate, color, form, surface characteristics, and optical nature. This method is most effective to identify the source from trees' tissue and pollen. But for the analysis by this method, a large number of particles must be observed, and it takes time and cost. It has also the disadvantage that its sensitivity is not sufficient to identify the organic particles and non-crystalline particles.

#### (2) Physical analysis

#### (1) Trajectory line analysis

This method is to identify the source by following up the trajectory line from the monitoring point to windward in order, based on the data of wind direction and velocity.

This method is often used to identify the source when the concentration of pollutants has reached a high level. This method is rather simple but the contribution of a certain source cannot be evaluated quantitatively.

#### (2) X-ray analysis

This method is to identify the crystalline particle quantitatively, such as mineral, cement dust, calcium carbonate, etc. But it is not suitable for non-crystalline particle and fine particles.

# (3) Chemical analysis

(1) Radioisotope method

This method is based on the principle that <sup>14</sup>C/<sup>12</sup>C of carbon compounds are measured and it is used for the apportionment of carbon from crude oil and other carbon. The half life of <sup>14</sup>C is 5,730 years and so the crude oil is identified as not containing <sup>14</sup>C because it has been stored and accumulated for several hundred million years. <sup>1)</sup>

(2) Concentration factor method

This method is to measure the concentration factor of the element (i) contained in the particles and to evaluate the impact by the emission sources, as shown in the following equation. If the data for chemical components at the emission sources are available, the quantitative analysis is also possible.

$$EFi = \frac{(Ci/Cs) \text{ aerosol}}{(Ci/Cs) \text{ reference}}$$

where; the concentration ratio is between trace element (i) and standard element (s).

(3) Time series analysis and space distribution analysis

These methods are to assume the emission source from the time series correlation of particulate weight and chemical components, and from space distribution of chemical components and those at the emission sources. The method itself is rather simple, but it is not suitable to identify the particular sources and quantitative assumption.

(4) Chemical Mass Balance method

This method was proposed by Miller et al.<sup>19)</sup> and it is to estimate the emission sources quantitatively by statistical processing of chemical components of the particulates at the monitoring point and emission source.

(5) Multivariate analysis

Multivariate analysis methods are main components analysis, factor analysis, regression analysis, cluster analysis, etc. These methods are all to obtain the information related to the emission sources from the variation between many samples collected for chemical components of the particulates. For these methods, advance information on the type of sources and chemical components are not required. These methods have been further improved into the TTFA method (Target Transformation Factor Analysis)<sup>20)</sup> which is to estimate the emission sources contribution rate quantitatively. This TTFA method has been recently highlighted among researchers.

# 2.2 Estimation of Contribution Rates of Emission Source Types on Particulate Matters by CMB Method

In this chapter, using chemical components data measured during each short term field survey, we estimate the contribution rates of each emission source type on particulate matters.

#### 2.2.1 Outline of the CMB (Chemical Mass Balance) Method

The CMB method is the technique that estimates the contribution rate of each emission source using two types of data: the measurement data of the chemical components on particulate matters at one receptor site during a certain period of time, and the data of the chemical composition of each emission source type. Recently, this method has received attention with the progress of analysis techniques. But, this method cannot distinguish the contribution rates of the similar emission source types, and the estimations are confined to the area of measurement stations and the prediction for future is impossible.

(1) Estimation method for contribution rates of source types on particulate matters by the CMB method

The following equation explains the basis of the CMB method.

An example showing the basis of the CMB method is given as follows: In this case, it is assumed that the emission sources are two types, and Al and Ca are emitted from only two source types: Soil and Factories.

The chemical components composition of the emission source on particulate matters is supposed as follows:

Emission Chemical sources components	Soil	Factorics
Al	10%	5%
Ca	1%	3%

The chemical concentration at a receptor site is supposed as follows:

Chemical concentration	10 μg/m³
Al	$0.55  \mu \mathrm{g/m^3}$
Ca	$0.08 \ \mu {\rm g/m^3}$

So the following matrices are defined.

$$C = \begin{bmatrix} C_1 \\ C_2 \\ \vdots \\ C_i \end{bmatrix} \quad F = \quad \begin{bmatrix} f_{11} & f_{12} & \dots & f_{1j} \\ f_{21} & f_{22} & \dots & f_{2j} \\ \vdots & \vdots & & \vdots \\ f_{i1} & f_{i2} & \dots & f_{ij} \end{bmatrix} \quad S = \quad \begin{bmatrix} S_1 \\ \vdots \\ S_j \end{bmatrix}$$

C<sub>i</sub>: The concentration of component (i) in suspended particulate matter at one receptor site

f<sub>ij</sub>: The proportion of component (i) concentration emitted from the source (j)

S<sub>i</sub>: The concentration of dust emitted from source (j)

j=1: Soil

j=2: Factories

In this case, the emission source types related to Al and Ca are restricted to Soil and Factories. Then, concerning Al, the concentration on particulate matter at one receptor site is the sum of two values; the product of Soil contribution rates and Al concentration rates in Soil, and the products of Factories contribution rates and Al concentration rates in Factories. The variables,  $S_1$  and  $S_2$  satisfy the following equations (2-2).

about Al 
$$0.10 \times S_1 + 0.05 \times S_2 = 0.55$$
about Ca 
$$0.01 \times S_1 + 0.03 \times S_2 = 0.08$$
 (2-2)

The following determinant (2-3) satisfies this simultaneous equation.

$$\begin{bmatrix} 0.55 \ \mu g/m^3 \\ 0.08 \ \mu g/m^3 \end{bmatrix} = \begin{bmatrix} 0.10 & 0.05 \\ 0.01 & 0.03 \end{bmatrix} \begin{bmatrix} S_1 & \mu g/m^3 \\ S_2 & \mu g/m^3 \end{bmatrix} \dots (2-3)$$

That is

The results are as follows:

$$S_1 = 5.0 \ \mu g/m^3$$
  
 $S_2 = 1.0 \ \mu g/m^3$ 

Then the contribution (rates) of Soil is 5.0  $\mu$ g/m<sup>3</sup> (50%) and that of Factories is 1.0  $\mu$ g/m<sup>3</sup> (10%).

In general, the chemical composition is determined for each emission source type (Soil, Sea salt particle, Fuel combustion, Refuse incincration, Automobile, etc.), and the contribution rates of each emission source type is estimated with the following equation (2-5):

$$C_i = \sum_{j=1}^{m} F_{ij} \cdot \alpha_{ij} \cdot S_j \qquad (2-5)$$

where;

 $a_{ij}$ : Conversion rate of component (i) from the source (j)

In general, the value of  $\alpha_{ij}$  is assumed to be 1, then Equation (2-5) is written as follows:

$$C_i = \sum_{j=1}^{m} F_{ij} \cdot S_j \tag{2-6}$$

The CMB method is the technique by which  $S_i$  is estimated using the known values,  $F_{ij}$  and  $C_i$ . When we measure the component concentration,  $C_i$  (i=1, 2, ... n), the following simultaneous equation can be derived from Equation (2-6).

$$C_{1} = F_{11} \cdot S_{1} + F_{12} \cdot S_{2} + \dots + F_{1m} \cdot S_{m} 
C_{2} = F_{21} \cdot S_{1} + F_{22} \cdot S_{2} + \dots + F_{2m} \cdot S_{m} 
C_{n} = F_{n1} \cdot S_{1} + F_{n2} \cdot S_{2} + \dots + F_{nm} \cdot S_{m}$$
(2-7)

where:

m: Number of emission sources

n: Number of receptor sites

Now we express Equation (2-7) in matrix, then

$$C=F\cdot S \qquad (2-8)$$

is obtained.

The methods of solution on Equation (2-6) have been variously proposed. Some of them are as follows:

- (1) Simultaneous equation method
- (2) Least-squares method
- (3) Weighted least-squares method
- (4) Effective variance method
- (5) Ridge regression method

#### 1) CMB method using simultaneous equation

If the number of  $C_i$  and  $F_{ij}$  is equal to that of emission sources in Equation (2-6), we can get the unique solution. So in this method we select market elements whose number equals to the number of emission sources and obtain the simultaneous equation. (Equation 2-6) Solving this equation, we estimate the contributions on the concentrations at receptor sites from emission sources.

When n equals to m, matrix F becomes a square matrix. Multiplying both sides of Equation (2-8) by  $F^{-1}$  (inverse matrix), we can get S.

$$S=F^{-1}\cdot C$$
 (2-10)

## 2) CMB method using least-squares method

If n (number of receptors) is greater than m (number of emission sources), Equation (2-6) generally has solutions S. In this method, we select  $\hat{S}$  (the presumed value of S) which makes the residual of C and  $\hat{C}$  ( $\hat{C}=F\cdot\hat{S}$ ) minimum. When we define the residual vector  $E=\hat{C}-C$ , sum of squares about residual  $\varepsilon^2$  is expressed by following equation:

$$\varepsilon^{2} = E^{t} \cdot E = (\hat{C} - C)^{t} \cdot (\hat{C} - C)$$

$$= (F \cdot \hat{S} - C)^{t} \cdot (F \cdot \hat{S} - C) \qquad (2-11)$$

where

$$(\hat{C}-C)^t$$
 is the transposed matrix of  $(\hat{C}-C)$ 

The condition which makes  $\varepsilon^2$  minimum is as follows:

$$\frac{\partial \epsilon^{2}}{\partial S} = 0$$

$$\frac{\partial \epsilon^{2}}{\partial S} = 2F_{t} \cdot (F \cdot \hat{S} - C) = 0$$
(2-12)

We transform Equation (2-12), then following equation is obtained.

$$\hat{S} = (F^{t} \cdot F)^{-1} \cdot F^{t} \cdot C^{-1} \cdot A \qquad (2-13)$$

# 3) CMB method using weighted least-squares method

Generally in the least-squares method, the whole data are treated equally. But it is usual for the accuracy of the data to differ. So we weigh the more accurate data with more weight and weigh the less accurate data with less weight. Such technique is called the weighted least-squares method. In this case, we adopted the inverse of the standard error (the standard deviation of accidental error) as the weight.

When we perform measurements of concentration  $C_i$  in k times, the value of  $C_i$  can be obtained as the following equation:

$$C_{i} = \frac{1}{k} \sum_{\alpha=1}^{k} C_{i\alpha} \qquad (2-14)$$

 $\sigma_i$  (the standard deviation of  $C_i$ ) and  $S_i$  (standard error) are defined as follows:

$$\sigma_{i} = \sqrt{\sum_{\alpha=1}^{k} (C_{i\alpha} - C_{i})^{2} / (k-1)}$$
 (2-15)

$$S_{i} = \frac{\sigma_{i}}{\sqrt{k}} = \sqrt{\sum_{\alpha=1}^{k} (C_{i\alpha} - C_{i})^{2} / k(k-1)}$$
 (2-16)

Thus, the basic Equation (2-6) is modified.

$$\frac{C_{i}}{S_{i}} = \frac{1}{S_{i}} \sum_{j=1}^{m} F_{ij} \cdot S_{j} = \sum_{j=1}^{m} \frac{F_{ij}}{S_{i}} \cdot S_{j}$$
 (2-17)

Here, we define the weight matrix W as follows:

$$W = \{W_{ij}^{2}\}\$$

$$W_{ij} = \left\{\frac{1}{S_{i}} (i=j) \atop 0 \ (i \neq j)\right\}$$
(2-18)

Using Equation (2-18), Equation (2-17) is expressed by matrix.

$$W^{1/2} \cdot C = W^{1/2} \cdot F \cdot S$$
 (2-19)

Then we can obtain the estimated value S

$$\hat{S} = \{ (W^{1/2} \cdot F)^t \cdot (W^{1/2} \cdot F) \}^{-1} \cdot (W^{1/2} \cdot F)^t \cdot W^{1/2} \cdot C 
= (F^t \cdot W \cdot F)^{-1} \cdot F^t \cdot W \cdot C$$
(2-20)

# (2) The emission source types and the composition of the emission sources

1) The emission source types

To estimate the contribution rates of the emission source types on particulate matters by the CMB method, it is necessary to determine the particular chemical components (the marker elements) of each emission source type. And it is difficult to estimate the contribution rate of the emission source type without marker elements.

At present, the marker elements are considered for these emission source types: Soil, Sea salt, Gasoline automobiles, Fuel oil combustion, Iron and steel industry, Refuse incineration, Cement (includes Road dust), etc.

2) The composition of the emission sources

To estimate the contribution rates in the particulate matters by the CMB method accurately, it is very important to take the composition data of each emission source type. But the composition data are variable depending on the conditions (operation and facilities), and so it is very difficult and takes a lot of money and time to measure the representative accurate data of the emission sources. And in the case of making use of the data according to the literature, the results of the estimation are not reliable because of the gap between this data and the actual emission sources.

3) The weight coefficient

To estimate the contribution rates on the particulate matter by the weighted least-squares method, it is required to calculate the error variance of each chemical concentration. In recent studies, someone used unique error variances (Watson<sup>21)</sup>), but Scheff<sup>22)</sup> decides them using the results of repeated analyses on samples. And Kowalczyk<sup>23)</sup> considered the variance of analysis error by instrument neutron activation analysis and the value of the filter blank.

# 2.2.2 The Estimation of the Contribution Rates of the Emission Source Types on Particulate Matters by the CMB Method

#### (1) Method of the estimation

According to the former discussion about the stability of the estimation results by the Industrial Pollution Control Association of Japan (IPCAJ)<sup>24)</sup>, the weighted least-squares method is adapted for the estimation.

#### (2) Emission source types

The contributions of the emission source types are estimated by using 8 emission source types. These are:

- (1)Sea salt
- (2) Soil
- (3) Diesel automobile

- (4) Gasolien automobile
- (5) Iron and steel industry
- (6) Fuel oil combustion
- (7) Refuse incineration
- ® Road dust

# (3) The composition of the emission sources

The chemical components of each emission source type are determined as follows:

(1) Seal salt		Na, Br
② Soil		Al, Ca, Sc, Fe, Ti
3 Diesel automobile		Elemental carbon, Organic carbon
(4) Gasoline automobile		Pb, Br
(5) Iron and steel industry		Mn, Fe, Zn, Cr
6 Fuel oil combustion		V, Ni
7 Refuse incineration		K, Zn, Sb
Road dust	• • • • • • • • • • • • • • • • • • • •	Pb, Ca, Fe, Sc

Table 2-1 shows the chemical compositions of the emission source types (emission source matrices). Now, the matrices of Soil and Road dust are determined by the actual samples collected in the Samut Prakarn industrial district, and other matrices are selected from the literature.

The grounds for setting the emission source matrices are explained as follows:

1) Sea salt particle

The chemical composition of the emission source is the same as the composition of the sea water itself, and the matrix data by Mizobata<sup>25)</sup> were cited.

(2) Soil and Road dust

The actual sample data of the soil and the road dust were used. The sampling points are illustrated in Fig. 2-2.

Table 2-1 The Matrix on the Component Concentrations in Emission Sources (unit; ppg)

	Sea salt	Soil	Diesel automobile	Gasoline automobile	Iron and steel Ind.	Fuel oil combustion	Refuse in- cineration	Road dust
Ai	0.3	55500	240	1300	18000	14250	9770	50000
Br	1900	17.5	0	20000	140	60	6420	150
Ca	12000	1750	3260	0	47000	14500	13600	26000
Cr	0.001	27.5	46	21	4230	2760	570	33
Fe	0.29	12000	30700	4900	301000	28300	5900	18000
K	11000	31500	120	2000	11500	1300	100000	21000
Mn	0.058	365	77	51	27700	500	540	610
Ka	304200	6500	120	200	7150	37800	88300	7200
Ni	0.014	30	0	38	1700	21000	135	25
Pb	0.087	70	490	120000	10500	900	53400	130
Sb	0.014	1.575	0.48	1.9	90	10	1120	3,4
Sc	0.001	4.45	0	0.12	2.9	0.33	0.74	4.5
Se	0.12	9.2	3.0	0	51	67	33	6.1
Ťi	0.029	2700	0 .	0	1000	430	1380	2300
٧	0.058	35.5	2.9	2.1	365	37700	27	46
Zn	0.029	17.5	1160	1400	26000	1700	79400	130
Cele	0	3250	620000	301000	_33000	371000	94000	16900
Corg	0	1000	100000	263000	33000	24000	15000	600

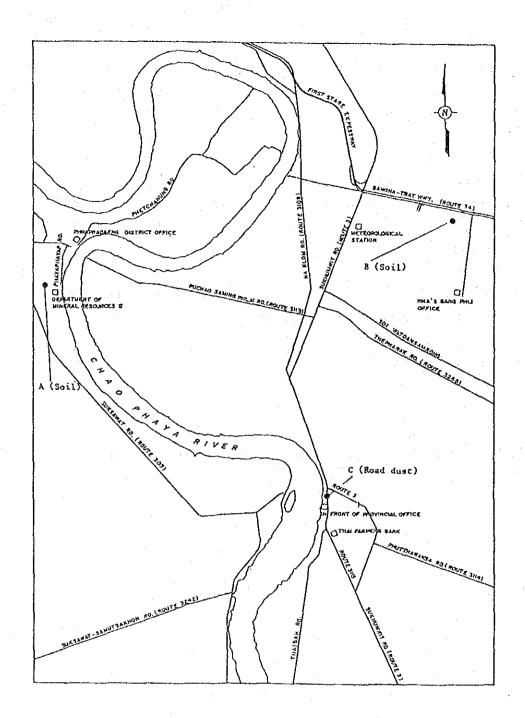


Fig. 2-2 Location of the Sampling Points for Soil and Road Dust

Soil: used the average data of these sampled at points A and B

Road dust: used the road dust data sampled at point C

(3) Diesel automobile

The average of the data of Fukuzaki<sup>26)</sup>, Hayashi<sup>27)</sup> and Koyama<sup>28)</sup> was used.

(4) Gasoline automobile

The data of the leased gasoline automobile by Mizobata<sup>25)</sup> was used. Moreover, the data of the carbon by Naito<sup>29)</sup> was added.

(5) Iron and steel industry

The average of the data of Mizobata<sup>25)</sup>, Catz<sup>30)</sup>, and Scheff<sup>31)</sup> was used. And the data of the carbon by Hayashi<sup>27)</sup> was cited.

(6) Fuel oil combustion

The average of the data of Mizobata<sup>25)</sup>, Hayashi<sup>27)</sup>, Scheff<sup>31)</sup>, Friedlander<sup>32)</sup> and Kowalczyk<sup>33)</sup> was used.

(7) Refuse incineration

The average of the data of Mizobata<sup>25)</sup> and Scheff<sup>31)</sup> was used.

(4) The chemical concentrations on particulate matters (The chemical concentrations at the receptor sites)

The chemical components data used for the estimation of the contribution rates of the emission source types on particulate matters were measured by the low-volume sampler at each monitoring station during each short time field survey. The method of the analysis and the data of these chemical components (33 elements, 7 ions, elemental carbon and organic carbon) were as shown in PART II. Then 18 chemical components shown in Table 2-2 are selected as the index elements. In addition, the contribution rates are not estimated using the data of the Andersen sampler, because the carbon components data are lacking.

Table 2-2 Monitoring Data of Chemical Component by Low Volume Sampler for the CMB Method (unit: ng/m³)

Chemic		1 s 1	Su	гчеу			2 n	i Su	rvey			3 r c	1 Su	rvey	
al Com ponent	MS I	MS 2	KS 3	MS 4	MS 5	HS 1	MS 2	MS 3	MS 4	KS 5	NS I	MS 2	MS 3	MS 4	XS 5
Al	1300	810	1100	1400	2200	760	310	490	650	570	770	410	410	820	740
Вг	42	17	31	37	18	14	9	19	11	12	26	4.5	9.5	15	7.2
Ca	2100	1200	2000	1600	2100	730	<800	2000	1600	<500	1700	800	1300	1500	1300
Cr	6.5	7.3	31	5.1	3.3	3.7	10.0	10.0	2.9	0.77	7.9	7.9	7.3	4.5	5.0
Fe	1700	1300	4700	1700	1100	700	420	2000	400	260	1200	840	1700	940	450
K	1300	1100	1700	1200	2000	640	350	1300	750	540	<800	1400	1100	<800	<900
Mn	60	58	360	75	41	50	11	200	24	15	71	23	72	34	21
Na	1100	1700	1400	1500	1100	1400	2800	2100	1900	1800	780	1000	840	1100	790
Ni	19	< 20	47	< 10	< 8	< 7	19	55	6	< 6	< 10	22	18	< 10	15
Pb	450	1590	780	290	30	50	20	380	50	< 10	220	1100	230	140	40
Sb	29	64	17	17	5.4	5.1	1.4	6.5	1.7	0.47	5.6	32	10	4.1	3.1
Sc	0.20	0.13	0.19	0.20	0.23	0.11	0.054	0.078	0.076	0.071	0.15	0.11	0.10	0.14	0.15
Se	<0.9	<1.0	<2.0	<0.8	0.96	1.1	51	1.5	0.5	0.95	1.7	<1.0	8.0>	<0.7	<0.6
Ti	67	86	130	- 140	150	97	93	57	<100	82	62	< 40	66	59	51
V	12	16	35	7.4	5.0	6.4	26	63	4.5	2,1	8.7	13	22	6.3	3.4
Zn	720	-900	3700	760	140	250	82	2900	100	18	960	370	1100	370	160
Cele	15300	13500	16400	16200	9500	7500	7500	11500	11500	6200 .	10600	5500	8100	9700	4000
Corg	8700	5800	10800	9400	5700	4300	4000	5300	5800	4100	2700	2000	3200	4000	2000

# (5) The weighted coefficient

The weighted coefficient is the reciprocal of the measurement error. The chemical components shown in Table 2-2 were measured by the neutron activation analysis except for Pb and elemental carbon and organic carbon; Pb was measured by the X-ray fluorescence and these carbons were measured by the thermal analysis. So, now the precision (the rates of variance on four measurement data) of the chemical components by IPCAJ<sup>24</sup>) are mainly adapted, because the method of analysis is similar to this preceding report. The precisions of the components (K, Ni, Se) of which measurement errors were not evaluated because of the lack of effective data, and of the components (elemental carbon & organic carbon) of which measurement errors were not reported were cited from the reports of Nagatsuka<sup>34</sup>) and Hayashi<sup>35</sup>), respectively. These precisions are shown in Table 2-3.

In addition, the concentrations of the components under the detection limit shown in Table 2-2 were assumed to be 1/2 times of detection limits and the precisions are assumed 10 times (1000%) values.

Table 2-3 Precision of Chemical Component (unit: %)

Compo- nent	IPCAJ	Naga- tsuka et al.	Hayashi	Adopted
Al	5.8	15.7		5.8
Br	_ 27.0	33.0		27.0
Ca	16.6	8.4		16.6
Cr	10.1	3.9		10.1
Fe	11.4	15.0	į	11.4
K		5.6	-	5.6
Mn	10.1	7.8		10.1
lla	10.1	13.8		10.1
Ni		: ·		10.0
Pb	11.0		-	11.0
Sb	5.6	22.6		5.6
Sc_	9.6	22.9		9.6
Se	—			10.0
Ti	39.4			39.4
Y	5.8			5.8
Zn	7.4	10.7	_	7.4
Cele	1.7		1.7	1.7
Corg	4.8		4.8	4.8

# 2.2.3 The Results of the Estimation by the CMB Method

# (1) The results of estimation by the CMB method with 8 emission source types

The contribution rates of the emission source tyes on particulate matters were estimated by the weighted least-squares method at every measurement station (MS1-MS5) during each short term field survey (1st-3rd). The results are shown in Table 2-4.

Table 2-4 The Results of the Estimation by CMB Method (8 Emission Sources)

[First survey]					(%)
Component	MS 1	MS 2	MS 3	HS 4	MS 5
Sea salt	-0,2	2.6	-0.7	1.7	3,9
Soil	-65.1	-40.7	-30.8	-59.7	57.1
Diesel automobile	35.0	34.8	27.4	32.8	27.1
Gasoline automobile	-0.3	-3.4	-2.2	-2.0	-1.3
Iron and steel Ind.	0.0	0.7	8.9	-0.5	0.2
Fuel oil condustion	0.4	0.6	1.0	0.2	0.1
Refuse incineration	15.9	17.2	16.7	13.0	3.7
Road dust	112.5	68.5	53.1	102.7	29.6
TOTAL	98.2	80.4	73.4	88.2	120.3

[Second survey]					(%)
Component	HS 1	MS 2	MS 3	MS 4	MS 5
Sea salt Soil Diesel automobile Gasoline automobile Iron and steel Ind. Fuel oil combustion Refuse incineration Road dust	9.9 -15.7 34.3 -3.2 0.8 0.4 9.6 60.4	20.6 8.3 31.1 -0.9 0.6 1.9 2.6 5.9	9.2 53.7 36.3 -0.6 3.2 3.3 12.5 -43.0	15.3 87.7 49.6 -0.5 1.0 0.3 3.1 -64.5	22.6 36.8 35.2 -0.3 -0.2 0.2 0.7 3.6
TOTAL	96.6	70.1	74.6	92.0	98.5

				(%)
MS 1	MS-2	MS 3	MS 4	MS 5
1.4	4.2	-0.2	4.4	6.3
-109.0	57.8	-22.7	-67.5	-1.0
38.4	29.2	32.5	33.7	22.9
-0.3	-4.6	-5.2	-0.9	~2.1
2.4	2.7	2.4	0.8	1.8
0.4	1.1	1.4	0.3	0.3
11.8	17.5	21.9	7.8	7.2
154.6	-40.8	42.0	108.3	48.4
99.8	67.1	72.2	86.9	83.8
	1.4 -109.0 38.4 -0.3 2.4 0.4 11.8 154.6	1.4 4.2 -109.0 57.8 38.4 29.2 -0.3 -4.6 2.4 2.7 0.4 1.1 11.8 17.5 154.6 -40.8	1.4 4.2 -0.2 -109.0 57.8 -22.7 38.4 29.2 32.5 -0.3 -4.6 -5.2 2.4 2.7 2.4 0.4 1.1 1.4 11.8 17.5 21.9 154.6 -40.8 42.0	1.4     4.2     -0.2     4.4       -109.0     57.8     -22.7     -67.5       38.4     29.2     32.5     33.7       -0.3     -4.6     -5.2     -0.9       2.4     2.7     2.4     0.8       0.4     1.1     1.4     0.3       11.8     17.5     21.9     7.8       154.6     -40.8     42.0     108.3

From the results, the abnormalities such as negative contribution rates on Soil, the over 100-percent contribution rates on Road dust and so on are suggested. Furthermore, the correlation coefficients among each emission source type are calculated with the chemical composition data (Table 2-1), and the results are shown in Table 2-5.

The result that the correlation coefficient between Road dust and Soil is 0.866 indicates the both emission source types are very similar. In the case of the CMB method, the similarity of some emission source types brings about the accidental error called property of multiple collinearity, and so the estimation is not possible because of instability. In this case, it appears to be in this condition.

Table 2-5 The Correlation Matrix among Emission Sources

Emission source	Sea salt	Soil	Diesel	Gaso- line	Iron Steel	Fuel oil	Refuse	Road dust
Sea salt Soil Diesel automobile Gasoline automobile Iron and Steel Ind. Fuel oll combustion Refuse incineration Road dust	0.014 -0.077 -0.115 -0.079 0.015	1.000 -0.064 -0.136	-0.064 1.000 0.794 0.065 0.983 0.432		0.113 0.065 0.007 1.000 0.067 -0.068	-0.041 0.983 0.710 0.067 1.000 0.452	0.432 0.345 -0.068 0.452 1.000	0.866 0.153 -0.010

# (2) The results of the estimation by the CMB method with 7 emission source types

The results calculated by the CMB method with 8 emission source types were unstable due to the similarity of the component compositions between Soil and Road dust (the correlation coefficient is 0.866). So in the next place, the contribution rates are estimated with 7 emission source types; the average value of Soil is adapted as Soil+Road dust.

The results shown in Table 2-6 indicates that the contribution rates of Gasoline automobile are negative at many stations, and the contribution rates of Refuse incineration are abnormally high. And the total contribution rate is over 100 percent at MS5 (1st survey).

Table 2-6 The Results of the Estimation by the CMB Method (7 Sources)

Component	KS 1	NS 2	HS 3	NS 4	MS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel Ind. Fuel oil combustion Refuse incineration	1.4 32.2 37.3 1.3 1.0 0.5 12.1	4.1 20.5 36.2 -2.0 1.4 0.6 13.6	0.0 16.1 28.4 -1.4 9.3 1.0 15.1	3.0 27.8 35.0 -0.7 0.2 0.2 10.0	4.1 82.3 27.8 -1.3 0.3 0.2 3.6
TOTAL	85.7	74.3	68.5	75.5	116.9
[Second survey]			. •		(%)
Component	MS 1	MS 2	NS 3	HS 4	HS 5
Can call	10.0	νν ε	0.0	15.0	- 00 6

Component	MS 1	MS 2	NS 3	HS 4	HS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel Ind, Fuel oil combustion Refuse incineration	10.3 35.0 35.7 -2.9 1.2 0.4 8.9	20.5 13.3 31.2 -0.9 0.7 1.9 2.6	9.0 15.4 35.5 -0.9 2.9 3.3 12.9	15.2 32.2 47.8 -0.6 0.5 0.3 3.3	22.7 39.9 35.3 -0.3 -0.2 0.2 0.7
TOTAL	88.6	69.3	78.1	98.7	98.2

[Third survey]					(%)
Conponent	MS 1	MS 2	MS 3	NS 4	NS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel Ind. Puel oil combustion Refuse incineration	3.1 17.1 41.5 1.4 3.8 0.4 7.6	3.8 21.4 28.4 -5.0 2.3 1.1 18.0	0.3 14.9 33.3 -4.6 2.7 1.4 20.7	5.0 18.0 36.0 -0.1 1.7 0.3 6.0	6.7 37.7 24.0 -1.9 2.3 0.3 6.8
TOTAL	74.9	70.0	68.8	67.0	75.9

(3) The results of the estimation by CMB method with 7 emission source types (The emission source matrices of Gasoline automobile, Diesel automobile and Refuse incineration are modified)

The results of the estimation with 7 emission source types and abnormalities such as negative contribution rates and abnormally high contribution rates. This explains why the emission source matrices of Gasoline automobile, Diesel automobile and Refuse incineration are unsuitable, so the reexamination is made as follows:

# (1) Gasoline automobile

Ca concentration was 0 ppm in the adapted emission source matrix of Gasoline automobile. But it appears that the emission source of Gasoline automobile contains Ca. So, Ca concentration in the emission source matrix of Gasoline automobile is modified to 1800 ppm according to the data by Fukuzaki<sup>26)</sup>.

#### (2) Diesel automobile

The average values of the data by Fukuzaki, Hayashi<sup>27)</sup> and Koyama were the adapted emission source matrix of Diesel automobile. Table 2-7 shows the comparison of Fe concentrations by each literature.

As this table shows, the variance of these data is large and the data by Fukuzaki is higher than others. So the value of Fe concentration is set at 360 ppm according to Koyama.

Table 2-7 The Comparison of Fe Concentrations Included in the Emission Source (Diesel Automobile)

	Fukuzaki	Hayashi	Коуала
Fe	61000 ppa	0 ppm	360 ppa

#### (3) Refuse incineration

The average values of the data by Mizobata and Scheff were adapted to the emission source matrix of Refuse incineration. But there were large differences in both components composition data. In this case, the data by Mizobata is adapted for convenience. Table 2-8 shows the emission source matrices modified as before. So the contribution rates estimated with this modified emission source matrices are shown in Table 2-9.

Table 2-8 The Matrix on the Component Concentrations in Emission Sources (Gasoline Automobile, Diesel Automobile and Refuse Incineration are Modified)

	40.4						(unit;ppp)
	Sea salt	Soil +	Diesel	Gasoline	Iron and	Fuel oil	Refuse in-
	Sea Satt	Road dust	autozobi le	autozobile	steel ind.	condustion	cineration
Al	0.3	55500	240	1300	18000	14250	4200
Br	1900	17.5	0	20000	140	60	830
Ca	12000	1750	3260	1800	47000	14500	11000
Cr	0.001	27.5	46	21	4230	2760	850
Fe	0.29	12000	380	4900	301000	28300	5200
K	11000	31500	120	2000	11500	1300	200000
Mn :	0.058	365	77	51	27700	500	330
Na	304200	6500	120	200	7150	37800	120000
Ni	0.014	30	0	38	1700	21000	135
Pb	0.087	70	490	120000	10500	900	17000
Sb	0.014	1.575	0.48	1.9	90	10	610
Sc	0.001	4.45	0	0.12	2.9	0.33	0.46
Se	0.12	9.2	3.0	0	51	67	_33
Ti	0.029	2700	0	0	1000	430	900
V	0.058	35.5	2.9	2.1	365	37700	27
Žn	0.029	17.5	1160	1400	26000	1700	26000
Cele	0	3250	620000	301000	33000	371000	94000
Corg	0	1000	100000	263000	33000	24000	15000

Table 2-9 The Results of the Estimation by the CMB Method
(7 Sources; Gasoline Automobile, Diesel Automobile and Refuse Incineration are Modified)

Component	· NS 1	MS 2	K SM	MS 4	MS 5
Sea sait	2.5	6.0	0.9	4.7	1.9
Soil + Road dust	33.1	23.7	17.1	34.4	66.3
Diesel automobile	34.3	36.0	27.9	37.3	24.1
Gasoline automobile	5.2	1.9	3.2	3.2	-0.7
Iron and steel ind.	1.6	2.2	11.0	0.9	∹0.3
Fuel oil combustion	0.4	0.6	1.0	0.2	0.1
Refuse incineration	5.1	5.8	8.8	4.3	7.3
TOTAL	82.2	76.2	69.8	85.1	98.6

[Second survey]					(%)
Component	MS 1	MS 2	NS 3	MS 4	MS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel ind. Fuel oil combustion Refuse incineration	9.2 30.3 32.3 0.1 1.6 0.4 5.8	14.2 10.2 29.5 -0.3 1.0 1.8 3.4	5.2 12.0 27.3 2.4 2.2 2.7 10.7	12.6 28.8 47.7 0.0 0.8 0.3 5.6	20.0 38.8 33.7 -0.2 -0.3 0.2 1.7
TOTAL	79.7	59.9	62.6	95.9	93.9

[Third survey]	· <u> </u>				(%)
Component	MS 1	MS 2	MS 3	HS 4	NS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel ind. Fuel oil combustion Refuse incineration	4.5 23.5 38.1 3.8 4.4 0.4 1.4	-0.5 21.3 26.1 0.1 2.4 1.1 24.8	0.5 14.9 28.5 1.5 2.9 1.3 13.0	6.4 23.5 37.6 2.3 2.4 0.3 2.0	6.1 30.9 23.3 0.2 3.0 0.3 6.0
TOTAL	76.0	75.4	62.6	74.6	69.8

At many monitoring stations, the contribution rates of Gasoline automobile, Iron and steel industry and Sea salt are negative, and the contribution rate of Refuse incineration is abnormally large as the former estimation (2).

# (4) The results of the estimation by the CMB method with 6 emission source types

There is no large refuse incineration factory in Sumut Prakarn industrial district. In spite of this condition, the former results  $((1)\sim(3))$  were estimated with 7 or 8 emission source types including Refuse incineration. In Refuse incineration, a large amount of Pb, K and Na are contained and these components are also in Gasoline automobile, Iron and steel industries and Sea salt. So, it seems Refuse incineration brought about the confusing results. Then, in the next place, the contribution rates of emission source types are examined with 6 source types excluded Refuse incineration. Table 2-10 shows the results of this estimation.

These results seem to be proper except for the over 100 percent contribution rates at MS5 (first survey) and at MS4 (second survey).

Table 2-10 The Results of the Estimation by the CMB Method (6 Sources; Refuse Incineration are Excluded)

(First survey)					(%)
Component	MS 1	MS 2	MS 3	MS 4	MS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel Ind. Fuel oil combustion	4.6 41.1 34.8 5.6 2.2 0.4	9.0 29.4 36.9 1.9 2.9 0.6	4.6 23.9 28.9 3.5 12.7 1.0	6.5 42.8 37.8 3.6 1.5 0.2	4.5 75.7 24.7 0.2 0.9 0.1
TOTAL	8.88	80.6	74.6	92.4	106.1

[Second survey]					(%)
Component	HS 1	MS 2	MS 3	MS 4	MS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel Ind, Fuel oil combustion	12.2 41.5 32.8 0.8 2.4 0.3	21.6 14.2 29.9 0.2 1.1 1.8	10.9 15.1 28.7 2.8 4.3 2.7	15.7 34.6 48.4 0.7 1.6 0.3	21.6 39.2 34.3 0.0 0.2 0.1
TOTAL	90.1	68.8	64.5	101.3	95.4

[Third survey]					
Component	MS 1	MS 2	MS 3	MS 4	MS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel Ind. Fuel oil combustion	5.1 28.3 38.2 3.9 4.5 0.4	10.0 27.1 29.8 0.2 3.6 1.2	5.8 18.9 30.2 1.9 4.9 1.3	7.4 30.8 37.8 2.4 2.6 0.3	8.4 48.3 23.8 0.9 3.3 0.3
TOTAL	80.4	71.8	63.0	81.5	84.9

# (5) The results of the estimation by the CMB method with 7 emission source types

The results estimated with 6 emission source types (excluding Refuse incineration) seemed suitable. However, the Sumut Prakarn industrial district there are some glass factories. So it is necessary to estimate the contribution rate of emission source type Glass industry. The chemical component composition data of Glass industry is cited from lide et al. <sup>36)</sup> and the contribution rates of emission source types are estimated. The emission source matrices are shown in Table 2-11, and the results of the estimation are shown in Table 2-12.

Table 2-11 The Matrix about the Component Concentrations in Emission Sources (Refuse Incineration is Replaced by Glass Industry.)

(unit;pps)

	Sea salt	Soil +	Diesel	Gasoline	Iron and	Fuel oil	Glass
	oca sait	Road dust	autogobile	autocobile	steel ind.	combustion	industry
Al	0.3	55500	240	1300	18000	14250	26700
Br	1900	17.5	0	20000	140	60	1080
Ca	12000	1750	3260	1800	47000	14500	31700
Cr	0.001	27.5	46	21	4230	2760	. 5160
Fe	0.29	12000	360	4900	301000	28300	10200
K	11000	31500	120	2000	11500	1300	23100
Mn	0.058	365	77	51	27700	500	7910
Na	304200	6500	120	200	7150	37800	42300
Ni	0.014	30	0	38	1700	21000	8820
Рь	0.087	70	490	120000	10500	900	39000
Sb	0.014	1.575	0.48	1.9	90	10	1800
Sc	0.001	4.45	0	0.12	1.3	0.33	2
Se	0.12	9.2	3.0	0	48	67	683
Ti	0.029	2700	0	0	_ 1000	430	2400
γ	0.058	35.5	2.9	2.1	365	37700	830
Zn	0.029	17.5	1160	1400	26000	1700	40100
Cele	0	3250	620000	301000	_33000	371000	371000
Corg	0	1000	100000	263000	33000	24000	24000

Table 2-12 The Results of the Estimation by the CMB Method (7 Sources; Refuse Incineration is Replaced by Glass Industry)

[First survey]

(%)

Component	AS 1	HS 2	HS 3	HS 4	MS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel Ind. Fuel oil combustion Glass industry	4.5 41.0 34.8 5.6 2.1 0.4 0.1	8.9 29.3 36.8 1.9 2.7 0.6 0.1	4.5 23.8 28.5 3.5 10.7 0.9 1.0	6.5 43.1 37.8 3.6 1.8 0.2 -0.2	4.4 76.2 24.5 0.1 0.3 0.1 0.5
TOTAL	88.6	80.4	72.9	92.9	106.2

[Second survey]

(%)

Component	NS 1	MS 2	MS 3	HS 4	KS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel Ind. Fuel oil combustion Glass industry	12.2 41.6 32.7 0.8 2.0 0.3 0.4	22.6 13.7 29.0 -0.4 0.4 1.6 2.0	10.7 14.8 27.0 2.4 0.1 2.4 3.7	15.8 35.1 47.6 0.4 0.3 0.2 1.4	21.3 41.7 33.9 -0.2 -0.7 0.1 0.7
TOTAL	89.8	68.9	61.1	100.9	96.9

[Third survey]

(%)

Component	MS 1	HS 2	MS 3	HS 4	MS 5
Sea salt Soil + Road dust Diesel automobile Gasoline automobile Iron and steel Ind. Fuel oil combustion Glass industry	5.0 28.3 38.0 3.8 3.8 0.3 0.5	9.7 26.2 28.7 0.1 2.6 1.1 2.1	5.7 18.7 29.8 1.9 4.0 1.3 0.7	7.4 30.5 37.6 2.4 2.0 0.3 0.6	8.4 47.1 23.1 0.5 2.3 0.2 1.6
TOTAL	79.8	70.4	62.1	80.7	83.3

These results indicate that the contribution rates of Gasoline automobile (at two stations), Iron and steel industry (one station) and Glass industry (one station) are negative. And the total contribution rate is over 100% at MS5 (1st survey) and MS4 (2nd survey).

In the next place, the comparison between the calculated value and the observed value is made in order to examine the adaptability of the CMB method on each index element. The result is shown in Table 2-13.

The results that the ratios (the calculated values: the observed values) for Al, Na, V, Ni, Pb, elemental carbon and Mn are nearly 1 show the high precision of the contribution rates estimation of Soil, Sea salt, Diesel automobile, Gasoline automobile, Iron and steel industry and Fuel oil combustion. On the other hand, some ratios for Sb and Zn are low (about 0.1), so the contribution rate of Glass industry is seemed to be inaccurate.

The reason for the inferior results of this estimation—some total contribution rates are over 100%, some contribution rates are negative and the estimated values of some components are different from the observed values—seems to be in emission source matrices. This means that this estimation data by using literature departed from the actual condition at the Samut Prakarn industrial district. Anyway, some negative contribution rates can be regarded as zero, because their absolute values are very small. So these results seem to be proper.